

SUBMITTED VERSION

Xuan Thi Dan Huynh, Tien Dung Khong, Adam Loch, Huynh Viet Khai

Solid waste management program in developing countries: contingent valuation methodology versus choice experiment

Environment, Development and Sustainability: a multidisciplinary approach to the theory and practice of sustainable development, 2022; 1-23

© The Author(s), under exclusive licence to Springer Nature B.V. 2022

This is a pre-print of an article published in *Environment, Development and Sustainability: a multidisciplinary approach to the theory and practice of sustainable development*. The final authenticated version is available online at: <http://dx.doi.org/10.1007/s10668-022-02572-4>

PERMISSIONS

<https://www.springer.com/gp/open-access/publication-policies/self-archiving-policy>

Self-archiving for articles in subscription-based journals

Springer journals' [policy on preprint sharing](#).

By signing the Copyright Transfer Statement you still retain substantial rights, such as self-archiving:

*Author(s) are permitted to self-archive a **pre-print** and an author's accepted manuscript version of their Article.*

a. A pre-print is the author's version of the Article before peer-review has taken place ("Pre-Print"). Prior to acceptance for publication, Author(s) retain the right to make a Pre-Print of their Article available on any of the following: their own personal, self-maintained website; a legally compliant pre-print server such as but not limited to arXiv and bioRxiv. Once the Article has been published, the Author(s) should update the acknowledgement and provide a link to the definitive version on the publisher's website: "This is a pre-print of an article published in [insert journal title]. The final authenticated version is available online at: [https://doi.org/\[insert DOI\]](https://doi.org/[insert DOI])".

17 August 2022

<http://hdl.handle.net/2440/135986>

Environment, Development and Sustainability

Solid waste management program in developing countries: contingent valuation methodology versus choice experiment

--Manuscript Draft--

Manuscript Number:	ENVI-D-21-02878R1	
Full Title:	Solid waste management program in developing countries: contingent valuation methodology versus choice experiment	
Article Type:	Original paper	
Keywords:	Disposal, Mekong River Delta, source-separation, strategic actions, willing to pay	
Corresponding Author:	Khong Tien Dung Can Tho University Can Tho, Can Tho VIET NAM	
Corresponding Author Secondary Information:		
Corresponding Author's Institution:	Can Tho University	
Corresponding Author's Secondary Institution:		
First Author:	Khong Tien Dung	
First Author Secondary Information:		
Order of Authors:	Khong Tien Dung	
	Dan Xuan Thi Huynh, M.Sc	
	Adam Loch, PhD, Associate Professor	
	Viet Khai Huynh, PhD, Associate Professor	
Order of Authors Secondary Information:		
Funding Information:	Can Tho University Improvement Project (VN14-P6)	Dr. Khong Tien Dung
Abstract:	<p>The willingness-to-pay by households in Vietnam to improve collection and processing services linked to source-separated municipal solid waste (MSW) is investigated in this research. This study contributes to current literature by comparing welfare benefits derived from contingent valuation methodology (CVM) and choice experiment (CE) for changes to waste service provision. The respondents were also stratified, for the first time, by households with and without previously sorted waste and by urbanization grade. Our results return broadly consistent willingness-to-pay estimates across the two approaches, and offer evidence that MSW sustainable management—a priority need in developing countries—may be addressed by positively enticing residents into a new fee-for-service program, but where social benefits from such engagement are made clear. Interestingly, residents in lower urbanization grades and those who have not practiced sorted waste reveal higher WTP values. Our results suggest that respondents are most interested in a wider selection of separated material options (i.e., recycled, organic and residual) as well as the potential for CO₂ emission reductions associated with new waste management programs. We argue that an application of both stated preference techniques widens the set of policy input factors available to government officials and useful evidence for structuring future engagement programs to address those who retain a preference for the status quo.</p>	
Response to Reviewers:	<p>Response Letter - Ms. Ref. No.: ENVI-D-21-02878 Title: Solid waste management program in developing countries: contingent valuation methodology versus choice experiment</p> <p>Editor's comments: Based on the advice received, I feel that your manuscript could be reconsidered for publication should you be prepared to incorporate major revisions. When preparing</p>	

your revised manuscript, you are asked to carefully consider the reviewer comments which are attached, and submit a list of responses to the comments. You are also requested to highlight the changes made in the revised manuscript.

Response: We wish to thank the Editor and five anonymous reviewers for very helpful comments and valuable suggestions to improve our paper. We have carefully noted the Editor's and Reviewer's recommendations and responded to them in detail below.

Reviewers' comments:

Reviewer #3:

In general, the manuscript is well written, but typos and grammatical errors are still found in different parts of the paper. Several sentences seem to be incomplete and/or confusing/convoluted. Authors are highly advised to have a thorough language proofreading.

Response: We appreciate and thank the Reviewer for pointing that out. We have corrected any typos and grammatical errors and double-checked the sentence structures.

The authors need to acknowledge the rich body of literature on waste sorting and management subject beyond the developing country including recent studies applying stated preference methods to elicit household preferences for waste sorting (e.g. Consumers in a circular economy: Economic analysis of household waste sorting behavior. Ecological Economics, 166, p.106402.

<https://doi.org/10.1016/j.ecolecon.2019.106402>. Even though the authors focus on a developing country context, still the authors need to justify and demonstrate the novelty of the design of their stated preference (e.g. choice attributes) in comparison to the designs in previous studies that have been carried out elsewhere. In the Literature Review section, the authors do include references, but they are rather limited and the section lacks critical overview of evidence from existing literature to highlight for example on drivers/determinants of household waste sorting behaviour/preferences which would be critical to inform the design of their present study.

Response: We appreciate and thank the Reviewer for pointing that out. We agree that the original review was limited for brevity reasons, and are very happy to include more references to clarify our contribution with reference to more recent and expanded sets of household preferences for waste solutions and their drivers of support.

Lines 104 - 117: Lacking references. Are these authors' own assumptions? Clarify.

Response: We have included a citation for this from World Bank 2018.

Lines 188 - 189: Provide justification as to why the authors chose this particular approach (compared to others) for eliciting the respondents' response on payment/bid in the CVM.

Response: We have included justification for this from lines 190-192. This should be a closed-end question, not an open-end.

Lines 205 - 210: Provide a discussion on how these selected attributes compare with previous CE studies on waste sorting preferences.

Response: We have included a discussion for this to make clear with previous CE studies in the literature review section and highlighted,

Line 211-213: Clarify what the authors mean by survey packages and survey versions. How are the two related?

Response: We have updated the text on this paragraph to make it clear by using the same words as in the survey versions.

Lines 211 - 217: Clarify the number of choice sets that each household was presented with during the CE survey.

Response: We have updated the text on this paragraph to make it clear.

Lines 220 - 221: Were the respondents presented with CVM and CE on the same day, one after the other? This must be quite demanding/exhausting to the respondents.

Critical reflection from the authors in the paper regarding this matter is needed.

Response: We thank the Reviewer for pointing this out to us, and we have rewritten that section to make it more clear.

Lines 225 - 233: In the paper, provide characteristics that define each of the three urban grades.

Response: We have updated the text on this paragraph to make it clear in the data collection section.

Line 244: "Accordingly, when undertaking the of" Something missing here?

Response: This has now been revised.

Line 246: What does tk in equation (2) signify? Clarify this in the paper.

Response: We have updated the text on this paragraph to make it clear.

Line 262 - 263: The sentence is incomplete and convoluted. Rewrite.
Response: This has now been revised.

Lines 309 - 311: Many CE studies have found that this IID assumption did not hold. The authors need to examine and demonstrate in the paper whether in their case the IID assumption holds. Otherwise, the authors should present alternative models that better capture preference heterogeneity (RPL, LCM).
Response: We thank the Reviewer for making this important point. We agree, and have employed the Hausman and McFadden (1984) test to check whether the IIA property is violated. We provide more discussion about this from Eq. 13 to Eq. 16 and add Table 5 as the MNL model result and Table 6 IIA test result.

Lines 356 - 478: These are mostly results. Discussion of the findings in relation to existing literature appears to be very limited. Major improvement in this section is inevitable.
Response: We have rewritten and expanded our discussion of the existing literature to ensure that this is now clearer in the Discussion section.

Lines 357 - 367: The authors only present the characteristics of their sample respondents but do not provide evaluation on how representative the sample is in relation to the target population. This is important to acknowledge and has important implications on the generalisability of the findings especially the fact that authors emphasize practical applications to local policy.
Response: We agree with the Reviewer, and have added the text to explain the representation of the sample by comparing it to the Vietnam Households Living standard survey by General Statistic Office.

Line 376 - 377: Provide the number (proportion).
Response: This has now been added.

Lines 513 - 518: Provide more concrete examples/proposals.
Response: This has now been added.

Reviewer #4:

In the paragraph from lines 117 to 122, on page 5, the author(s) mention the factors driving behavioral change in waste separation by private households. There is already statistical evidence on the factors that influence such behaviors in relation to the separation of solid waste by households at the urban level. I would particularly recommend reading and referencing the document:

Padilla, A. J., & Trujillo, J. C. (2018). Waste disposal and households' heterogeneity. Identifying factors shaping attitudes towards source-separated recycling in Bogotá, Colombia. *Waste Management*, 74, 16-33.

Response: This has now been added.

It would be convenient for the author(s) to specify which software they used for their experimental design. In addition, it would also be convenient for the authors to specify at the end which of the two methodologies is the optimal one in terms of policy design, considering the "high" percentage of opt-outs.

Response: This is a good point, and that detail has now been added.

The author(s) should justify why they choose a logit instead of a probit, and check the estimates without the opt-outs in a second estimation through a tobit model to corroborate the robustness of the design. In both methodologies, the author(s) also do not take into account the illegal disposal of waste, which is a key factor within MSW in developing countries.

Response: Again, we agree. We have rewritten and expanded our discussion of the econometric approaches to ensure that this is now clearer in the relevant section.

Reviewer #5:

The author(s) attempt to investigate the willingness-to-pay for solid waste in Vietnam using the CVM and CE valuation methods. In as much as I commend the authors for the effort, I have some concerns. I have outlined my comments for the necessary action before the paper can be considered.

In line 32. Your first statement of claim is not supported with facts and figures or even a reference. What is the evidence that Vietnam in developing countries in general are grappling with solid waste management? Kindly do the needful.

Response: A reference in support of this view has now been added.

Line 38: You are assuming that the problem is revenue so willingness-to-pay from the public sector may enable investments. On what premise is this based? Why is the problem not attitudinal? I think this has to be motivated a little bit.

Response: We agree with the Reviewer, and have included more discussion in the relevant section.

Again, why should the public pay for a private problem? Your argument for public investment does not present a paradigm shift in the narratives hence the problem will persist if the government cannot pay.

Response: The frame in this case is one of private payments to support the existing public investment that is insufficient to cover current and expected future funding requirements. Hence, our exploration of the level to which private payments may be elicited, set and ideally recovered. As this may not be as clear as we had hoped, we have rewritten where appropriate in an effort to address that.

In your Introduction, you have not been able to convince the reader regarding the reason why both approaches are important for solid waste management. Again, you have not been able to convince the reader why WTP is even important while public investment is being considered. You could have re-written your introduction using the study site to motivate the work better. Nonetheless, the current approach is not bad provided the study is well motivated.

Response: We agree with the Reviewer, and have included more citations from related previous research in literature in the introduction.

Line 104: It is unclear the fraction of the solid waste generated in Vietnam that is recycled.

Response: This sentence has now been revised.

Literature: Kindly show some related solid waste studies that compared methods and yielded similar results. You may want to add some few CVM or CE and other studies that yielded related estimates and why.

Response: We agree with the Reviewer, and have included more citations from related previous research in literature in the introduction.

The CVM hypothetical market must be clearly stated. I'm unsure how this was explicitly stated. Kindly quote the description.

Response: This has now been added.

Also, how the bidding format was applied is unclear. Yes, the stating bids have been given, how did it result in open-ended responses. How were the starting bids obtained from the managers and public officials with about VND30,000 difference? Why didn't you use actual market estimates for the various communities or areas?

Response: We have rewritten and expanded our discussion of the approaches to type of questions and bases for bid formulation to ensure that this is now clearer in the material section.

Why is the monthly fee for MSW management service for the CVM different from the CE. What informed this difference. Example \$0.86/month (CE) and \$0.87/month (CVM)

Response: We have rewritten and expanded our discussion of the approaches to ensure that this is now clearer in the material section.

Line 219: In the abstract, I see you used stratified sampling technique, however, in the method, I see just random sample. Which type of random sampling and how was this achieved?

Response: We appreciate and thank the Reviewer for pointing that out. We have corrected this as suggested.

In Table 1, is Option C the status quo? If yes, indicate it.

Response: This has now been added.

Line 282: Why is income controlled for as a dummy while the raw data obtained was not dummy? A Table of descriptive statistics even in appendix will do.

Response: That is a useful idea and has now been added. To follow up quickly here, the dummy income value is based on the amount of US\$385 issued by law in the National Assembly, and from which different levels of income were then stratified.

Line 357: The socio-economic characteristics are not compared with national data or estimates. Pls do!

Response: This has now been added and compared to the national survey.

Line 409: In Table 3, you present only the estimates without the margins or marginal effects. Pls do!

Response: We agree that the estimation of marginal effects in the Logit model is necessary in other cases. However, in the case of MWTP estimation, we only need the sign of coefficients to explain the relationship between the determinants and the probability of agreeing to pay for the SW management program. In addition, the coefficients generated from the Logit model must be used to estimate the MWTP instead of marginal effects. Therefore, we present only the coefficients in this case.

In Table 4, what is in parenthesis?

Response: This has now been added to the note for these numbers. I expect the authors to support their findings with existing literature.

Response: This has been raised by some of the other reviewers, and addressed as recommended.

The observations in Table 3 and 5 and different. Kindly justify in the paper.

Response: This has now been added in the material and methodology section related to the different approaches between CE and WTP.

Line 480: Why should a study with observations in Vietnam be interesting to all in developing countries

Response: We have expanded our discussion of this to ensure that this is now clearer in the material section.

Check how in-text citations with et al are presented. (XXXX et al., 2010) or XXXX et al. (2010)

Proofread the work thoroughly to correct minor typos

Response: Revised. Again, we are grateful to the Reviewer for their attention to detail.

Reviewer #6:

1- The English should be revised by a native English speaker.

Response: We thank the Reviewer for pointing that out. One of the authors is a native English speaker and we have used that person to carefully once more correct any typos and grammatical errors, and to double-check the sentence structures.

2- How to select the overall study sample as well as the number of samples in each city should be fully described.

Response: We have updated the text in this paragraph.

3- In the results section, the respondents' attitudes toward the MSW status are mentioned, while in this study, the respondents' attitudes were not evaluated. Based on the content, the respondents' experience in dealing with the issue seems to have been examined.

Response: We have updated the text in this in the material section.

4- The reason for the importance of the study and its implications, especially for global readers, has not been clearly and attractively stated.

Response: We have rewritten and expanded our discussion of the approaches to ensure that this is now clearer in the material section.

In total, I congratulate authors because it is an interesting paper. I suggest accept with minor revisions.

Response: We appreciate this positive feedback and thank the Reviewer for these opening comments about the value of the paper and its findings.

Reviewer #7:

The manuscript is to valuate the solid waste management programme in Vietnam using CVM and CE. Although the manuscript is well organised, there are some serious problems which need substantial revised to make it suitable for publication. The following are the detail comments:

1. Literature review: The results of similar studies should be reviewed and included in the literature review section. Currently the literature review focuses on methodologically review.

Response: We thank the Reviewer for suggesting this as it is a good idea and in line with other review comments. We totally agree, and this text has been added accordingly.

2. Methodology section: The questionnaire design has not been well described. I only see how the CVM and CE test are designed. None of the other data collected in this questionnaire survey has been included. As i aware in the results and discussion section, there are some other parameters used for statistical analysis. However, there are no information provided in the questionnaire design.

Response: The relevant sections of text have been revised in line with the Reviewer's suggestions.

3. The objective of this research is not clear. after reading the first part of the manuscript, it seems to me that this is a study in comparing two valuation methods and using solid waste management as an example for the evaluation of the two valuation method. However, i am aware that in the results and discussion section, this is not what I thought. It is a paper to evaluate the soild waster management programme in Vietnam. If so, there is a question needed to be answered, why use two valuation methods?

Response: We have rewritten and expanded our discussion of the approaches to ensure that this is now clearer in the material section.

4. if this manuscript is a study of the two valuation methods, there are nothing in the discussion sections about the good or bad of these methods.

Response: We do not view the work as a critique either way, as sufficient material is already in print on that subject. However, to address this for others, we have rewritten the paper where appropriate.

5. There is very limited in-depth discussion on the results with previous studies.

Whether your results are similar and difference with previous studies? and why?

Response: We thank the Reviewer for pointing this out to us, and we have rewritten it where appropriate in line with other suggested changes.

1 **Solid waste management program in developing countries: contingent valuation**
2 **methodology versus choice experiment**

3 Xuan Thi Dan Huynh¹, Tien Dung Khong¹, Adam Loch², Viet Khai Huynh¹

4 ¹School of Economics, Can Tho University, Vietnam

5 ²Centre for Global Food and Resources, School of Economics and Public Policy, Faculty of the
6 Professions, University of Adelaide, Adelaide South Australia, Australia

7 Corresponding author: Tien Dung Khong, Ph.D. (ktdung@ctu.edu.vn)

8

9

10

Declarations

11 **Funding:** This study is funded in part by the Can Tho University Improvement Project VN14-P6,
12 supported by a Japanese ODA loan.

13

Conflicts of interest/Competing interests:

- 14
- 15 • The authors have no conflicts of interest to declare that are relevant to the content of this
16 article.
 - 17 • The authors have no relevant financial or non-financial interests to disclose.
 - 18 • All authors certify that they have no affiliations with or involvement in any organization
19 or entity with any financial interest or non-financial interest in the subject matter or
20 materials discussed in this manuscript.
 - 21 • The authors have no financial or proprietary interests in any material discussed in this
22 article.

23

24 **Availability of data and material:** Data is available from the corresponding author upon request.

25

26 **Authors' contributions:** All authors contributed to the study conception and design. Material
27 preparation, data collection and analysis were performed by Xuan Thi Dan Huynh and Tien Dung
28 Khong. The first draft of the manuscript was written by Xuan Thi Dan Huynh and Tien Dung
29 Khong and all authors commented on previous versions of the manuscript. All authors read and
30 approved the final manuscript.

31

32

1 Solid waste management program in developing countries: contingent valuation 2 methodology versus choice experiment

3 4 5 6 7 4 **Highlights**

- 8 • Waste is a key issue for developing nations like Vietnam, with limited resources to cope.
- 9 • Social concern for environmental issues may lift households' WTP and increase resources.
- 10 • Identifying key attributes of preferred policy will strengthen MSW management and WTP.
- 11 • Identifying homogenous benefits may encourage those with status quo preferences to engage.

12 **Abstract**

13 The willingness-to-pay by households in Vietnam to improve collection and processing services
14 linked to source-separated municipal solid waste (MSW) is investigated in this research. This study
15 contributes to current literature by comparing welfare benefits derived from contingent valuation
16 methodology (CVM) and choice experiment (CE) for changes to waste service provision. The
17 respondents were also stratified, for the first time, **into** households with/without previously sorted
18 waste and **across** urbanization grades. Our results return broadly consistent willingness-to-pay
19 estimates across the two **methodologies**, and offer evidence that MSW sustainable management—
20 a priority need in developing countries—may be addressed by positively enticing residents into a
21 new fee-for-service program, but **only** where social benefits from such engagement are made clear.
22 Interestingly, residents in lower urbanization grades, and those who have not **previously sorted**
23 waste, reveal higher WTP values. Our results suggest that respondents are most interested in a
24 wider selection of separated material options (i.e., recycled, organic and residual) as well as the
25 potential for CO₂ emission reductions associated with new waste management programs. We argue
26 that an application of both stated preference techniques widens the set of policy input factors

27 available to government officials and useful evidence for structuring future engagement programs
28 to address those who retain a preference for the status quo.

29 **Keywords:** Disposal, Mekong River Delta, source-separation, strategic actions, willing to pay

30

31 **1. Introduction**

32 Governments worldwide are grappling with solid waste management, particularly in rapidly
33 developing countries like Vietnam. Unsuccessful waste management is associated with negative
34 environmental impacts including greenhouse gas emissions, land and water contamination, odour,
35 landscape deterioration and noise, and human health concerns such as fire hazard and disease
36 (Ghanbari et al. 2012). Effective waste management could reduce these negative impacts, but it
37 would most likely come at a high cost to society for improved infrastructure, modern
38 collection/sorting systems, and increased public awareness campaigns. Developing countries may
39 struggle to afford these increased costs, limiting their range of policy alternatives. However, if the
40 public is willing/able to pay for improved municipal solid waste (MSW) management the revenue
41 may enable investments to avoid amenity losses, human health risks, etc. (Pearce and Howarth
42 2000).

43 Reconciling the needs of waste producers with those of the waste managers is critical for
44 achieving policy objectives. To that end, stated preference techniques play an important role in
45 estimating both the willingness to pay (WTP) for a change to an environmental state and key
46 attributes for successful waste management policies (Ko et al. 2020). Waste management services
47 are frequently under- or non-priced, making it challenging to obtain economic benefit estimates
48 from market prices (Anaman and Jair 2000). Further, social preferences for waste services are not
49 homogenous across individuals and therefore policy-makers need to adopt targeted instruments
50 with appropriate incentives based on evidence (Massarutto et al. 2019). These characteristics also
51 support the selection and application of stated preference techniques—making them quite

52 common, as detailed below. Stated preference techniques also enable the benefit transfer of WTP
53 estimates beyond original study sites (Hanley et al. 1998a). This can be important for developing
54 countries if budget/time constraints make repeated valuation surveys at sites of interest impractical
55 (Damigos et al. 2016), and **where** a high level of precision is not required (Rosenberger and Loomis
56 2001).

57 **Household preferences for waste scheme management have been researched widely via**
58 **elicitation approaches such as choice experiments (Lee et. al, 2017) and CVM (Kipperberg and**
59 **Larson, 2012).** Yet studies of stated preferences for waste management in developing countries
60 remain limited (Ko et al. 2020). **This is** despite a recognition of the need for this work nearly two
61 decades ago, **especially via preference choice sets to** identify the attributes of effective waste
62 management systems (Jin et al. 2006), **and applications of discrete choice experiments aimed at**
63 **identifying the value provided by waste sorting services (Nainggolan et. al., 2019).** This motivates
64 our study of waste management preferences using both contingent valuation methods (CVM) and
65 choice experiment (CE) techniques to examine individual (private) drivers of willingness to pay
66 for a specific change to the environmental state, as well as the key (public) policy attributes that
67 might incentivize wider **adoption** by less-supportive individuals in the Mekong River Delta region
68 of Vietnam as a case study.

69 **2. Study site background**

70 The Mekong River Delta (MRD) has experienced a significant economic expansion in recent years
71 which has increased solid waste output. Migratory movement from rural to urban regions between
72 2009 and 2018 was caused by low and unstable **regional earnings**. Climate change has impacted
73 rice and other production in rural areas of the MRD, **while the** rapid development of industry and
74 services in urban areas of the six major MRD provinces/cities has created an imbalance in regional
75 income distribution. Rapid population expansion has depleted natural resources such as water,
76 electricity, and raw materials to fulfill production and consumption demands; negatively impacting

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

77 the environment. As a result, the amount of MSW created in metropolitan areas is rapidly
78 increasing.

79 MSW from urban areas accounts for more than half of all waste generated in Vietnam,
80 growing from 32,000 tons per day in 2014 to 37,000 tons per day in 2019 (Ministry of Natural
81 Resources and Environment of Vietnam, 2016; 2020). In 2020, the total **daily** volume of solid
82 waste created in MRD regions was around 14,000 tons. By contrast, the volume of solid waste
83 collected by government services each day is 9,800 tons or roughly 70% of the volume created
84 (Ministry of Natural Resources and Environment of Vietnam, 2020). Therefore, waste service
85 infrastructure, financial investment and human resources in the MRD region have not kept pace
86 with **output** growth. **Further, in** many MRD districts waste collection, transportation and treatment
87 operations are disconnected from one another, particularly at the municipal level where
88 public/private interaction is highest. Poor implementation of waste services is also **driving**
89 environmental and public health issues. In recent years environmental deterioration, particularly
90 in landfills, has been a significant source of concern. A program to improve municipal solid waste
91 (MSW) collection and treatment is therefore critical.

92 **An issue is the public cost of such change, and limited privately provided revenues.** The
93 present monthly rate for current MSW collection and treatment services of around US\$0.86 per
94 household is substantially lower than the service operating costs (Ministry of Natural Resources
95 and Environment of Vietnam, 2020), limiting government investment unless private individuals
96 are willing to pay more for the service in future. Increased pressure on government policy-makers
97 to identify an appropriate MSW management solution has come from a recent *National strategy*
98 *on Integrated Solid Waste Management: 2025 to 2050* approved by the Vietnamese Prime Minister
99 in Decision No. 2149/QD-TTg. The strategy sets specific objectives: **that,** by 2025, 90% of all
100 urban solid waste will be collected and treated according to national standards with 85% of all
101 MSW to be recycled, reused, energy-recovered or converted to organic fertilizer. This objective

102 necessitates the implementation of a more sophisticated waste management system with different
103 integrated solutions in the MRD. A focus of that change is the source separation of waste materials.

104 Source separation is the segregation of different types of solid waste (organics, plastics,
105 paper, non-recyclables) at the location **where** they are generated (household or business). The
106 number and types of categories into which wastes are divided usually depend on the collection
107 system used and their final destination/use (Moh, 2017). Advantages of source separation include

108 **(World Bank, 2018):**

109 (i) Economic: organic and recycled waste components in Vietnam account for about 60% - 65%
110 and 22% - 26% of solid waste, respectively. Organic components in residential solid waste
111 can be used to supply raw materials for fertilizer products in conjunction with the source-
112 separation of materials (e.g., composting bins or separate green waste collection). Recyclable
113 components such as plastic, glass, nylon, metal, and rubber can also be removed at source to
114 save money on incinerator operation costs while also increasing total landfill space for
115 residential solid waste.

116 (ii) Environmental: lowering the volume of organic and harmful materials (e.g., oils, paints and
117 chemicals) through source-separation reduces negative environmental effects such as smells,
118 leachate, and soil, groundwater and surface water contamination.

119 (iii) Social: Solid waste source separation also raises public awareness about environmental
120 conservation. Increased awareness, and the development of positive environmental habits,
121 can lead to increased social benefits.

122 In order to successfully accomplish a source separation initiative, private households must
123 be incentivized to change their behaviour (i.e., separate waste materials), and **contribute to** public
124 waste service **funding so that the government can update their** equipment, processes and employee
125 skills. As a result, we are interested in knowing what factors will encourage behavioural change
126 in the separation of waste, and if households would be willing to cover the additional costs of

127 upgraded services; where the fee paid in the future will obviously be higher than current, especially
128 in the case of developing country where this study will provide a useful base to compare with
129 other developing contexts (Padilla and Trujillo (2018)).

130 Both the requirement for increased revenue in support of changes to MSW management in
131 the MRD and the complex set of treatment criteria established under the new national strategy link
132 well with our adoption of stated preference techniques. Further, it is highly likely that there will
133 be a diverse set of preferences spread across the population spanning high acceptance of the need
134 for environmental change to those who prefer the status quo. Identifying these different groups
135 and their potential drivers is a key advantage of stated preference techniques. Finally, recent
136 reductions in oil prices highlight volatility in waste recycling markets and a need for governments
137 to be prepared to invest locally to address future instability (Ko et al. 2020). Our study of MSW
138 using stated preference techniques is therefore timely given a narrow focus on these issues in
139 developing countries since 2006. In support of that view, we begin with a review of the relevant
140 literature in this field.

141 3. Literature review

142 As stated above, solid waste management preferences are hard to assess since they often have no
143 market value and must be estimated using non-market valuation approaches. A willingness to pay
144 (WTP) for additional services that may change an environmental state is the underlying premise
145 of these approaches (Bateman et al., 2002). Optimal waste management systems ensure that
146 society gains a net maximum benefit from proper disposal (Garrod and Willis, 1998), and stated
147 preference techniques are commonly used to elicit people's WTP to establish those economic
148 benefits (Jin et al. 2006). The two most commonly applied techniques in the MSW literature are
149 contingent valuation methods (CVM) and choice experiments (CE). These techniques share a
150 common random utility theory basis allowing for direct comparison (bid) and are based on an

151 assumption that peoples' behaviour in a hypothetical market can indicate their genuine intentions
152 for environmental products (Hanley et al., 1998b), even with very little data (Diafas, 2016).

153 Both Damigos et al. (2016) and Ko et al. (2020) provide helpful summaries of applications
154 of CVM and CE techniques in the study of solid waste management. Both summaries indicate that
155 CVM techniques (see Aadland and Caplan 2006; Yusuf et al. 2007; Gillespie and Bennett 2013;
156 Ferreira and Maques 2015; Maimoun et al. 2016) are more commonly applied over CE—mainly
157 due to the complexities associated with CE design and implementation. However, despite these
158 issues, many studies employ CM to determine the MSW program characteristics for which people
159 are willing to pay (see Othman 2007; Adeoti and Obidi 2010; Czajkowski 2013; Yuan and Yabe
160 2015; Fukuda et al. 2018; Tarfasa and Brouwer 2018). In a seminal study, Adamowicz et al. (1998)
161 show that CE techniques may be beneficial over CVM in some contexts where they completely
162 explain the trade-offs between attribute qualities. CE techniques can also be used to eliminate bias
163 and other issues that might arise in surveys that include only “agree” or “disagree” questions
164 (Ready et al., 1996). Specific to MSW contexts, the capacity for CE to identify key attributes of
165 MSW programs that appeal to people is a necessary foundation for suggesting legislation or
166 program reforms to improve the efficacy of waste management services (Pearce et al., 2006).

167 By contrast, CVM surveys focus on recognizing the gains or losses from an environmental
168 change using a limited set of commodity features (Diamond and Hausman, 1994; Stevens et al.,
169 2000) where it may be appropriate to use an open-ended WTP question format. Where multiple
170 attributes are connected to an environmental change, applications of CVM may be inappropriate
171 (Jin et al. 2018). Further, in developing contexts where market distortions complicate shadow
172 pricing exercises (Arrow, 2001), and literacy rates are low, personal interviews using CVM
173 approaches may provide a more appropriate elicitation method (Johnston et al. 2017). Recognizing
174 the pros and cons of these techniques this study employs both CVM and CE approaches to i) assess
175 MRD household willingness to pay for improved solid waste management programs to reduce the

176 government's total financial burden in the Mekong Delta and ii) identify key policy attributes that
177 will inform future MSW policy design and implementation and increase engagement across a
178 wider sample of the MRD population. These are useful contributions in this growing literature,
179 where we will also compare estimated welfare differences—as suggested by Boxall et al. (1996)
—as important possible drivers of differences between the two modelled results.

181 There has been an increase in the number of studies using the CE method to evaluate various
182 attributes of sorted waste management programs (Das et al., 2008). Some studies, such as Othman
183 (2002), Jin et al (2006) and Karousakis and Birol (2008) have used CE methods to evaluate solid
184 waste management programs. However, it is necessary to clearly define the attributes of the waste
185 management program being studied. Past studies have often focused on the attributes associated
186 with the phases of a waste management system including: i) monetary attributes or the fee that
187 people pay for the improvement of the quality of services (Fukuda et al., 2018; Karousakis and
188 Birol, 2008; Othman, 2007; Sakata, 2007, Tarfasa and Brouwer, 2018); ii) the amount people may
189 receive for contributing to an improvement in the quality of solid waste management services
190 (Yuan and Yabe, 2015); iii) waste collection periods (Czajkowski et al., 2014; Fukuda et al., 2018;
191 Karousakis and Birol, 2008; Yuan and Yabe, 2015); iv) the availability of equipment to support
192 waste classification (Fukuda et al., 2018; Tarfasa and Brouwer, 2018; Yuan and Yabe, 2015) or
193 final waste treatment (Fukuda et al., 2018; Othman, 2007); and/or v) improved recycling methods
194 as an incentive to change behaviour (Sakata, 2007). However, in developing countries, the number
195 of solid waste classification types is often lower than that of studies conducted in developed
196 countries where solid waste might be classified into one of five (Czajkowski et al., 2014) or as
197 many as 11 categories (Sakata, 2007).

198 Therefore, when designing a CE survey the key attributes and levels must be carefully
199 identified and examined (Pearce et al., 2006), with the number of attributes proportional to the
200 number of observations in the study (Bateman et al., 2002); that is, the larger the number of

201 observations, the greater the number of attributes that can be included. However, to avoid complex
1
2 202 choice sets for respondents, the number of attributes should be limited to no more than four or five
3
4 203 attributes (Pearce et al., 2000; Hanley et al., 2002). Willis and Garrod (1999) argue that solid waste
5
6
7 204 management strategies should address public concerns about sustainable use of resources as well
8
9 205 as reduce the amount of solid waste that needs to be disposed of in landfills. This suggests four
10
11 206 attributes related to a household's MSW reduction program: reducing the amount of solid waste
12
13
14 207 that needs to be treated, reducing the amount of CO₂ emitted from solid waste treatment, the
15
16
17 208 number of types of solid waste classified, and the monthly fee for MSW management.

18
19 209 Further, while contemplating environmental action, psychological considerations should
20
21 210 also be taken into account. This broad term refers to an individual's personality and perceptual
22
23
24 211 qualities as they relate to their behaviour. Intrinsic incentives to engage such as behavioural
25
26 212 gratification (De Young, 1986) and threats to welfare from environmental problems (Baldassare
27
28
29 213 and Katz, 1992) are examples. A study should also consider any contextual factors, personal
30
31 214 capacities, attitudinal factors, and habitual factors (Soderholm, 2013) as key elements that
32
33
34 215 influence environmental behaviour. Finally, drivers of individual attitudes and motives for change
35
36 216 could stem from technical-organizational circumstances (external factors), socio-demographic
37
38
39 217 factors (e.g., age, gender, income), and socio-psychological variables (Miafodzyeva & Brandt,
40
41 218 2013).

44 219 **4 Research methodology**

46 220 *4.1 Survey instrument design*

48 221 For both the CVM and CE survey instruments **our** design process began with a thorough scan of
49
50
51 222 **the** available literature, and initial discussions/focus groups with both waste management experts
52
53 223 and local resident groups. This information was then used to establish requisite framing
54
55
56 224 information to be added **into** each survey, as well as **final** attribute and payment card details. In
57
58 225 each case, the penultimate survey designs were pre-tested with 20 randomly selected households
59
60
61
62
63
64
65

226 to calibrate the validity of the format, comprehension of the information included, and the
227 effectiveness of any bias management tools employed.

228 In the final iteration of the CVM survey, respondents were first introduced to the proposed
229 MSW program and told that it would aim to reduce landfill solid waste by supporting management
230 methods such as waste avoidance and reduction (e.g., recycling, reusing); where some further
231 detail about how this would work in practice via source separation was also offered. A cheap talk
232 script as an *ex-ante* bias correction was also shown to the respondents to remind them to consider
233 budget constraints and state an amount as if the payment was real. The respondent was then asked
234 if they would be willing to pay for the proposed program using closed-ended questions comprising
235 five possible monthly payment bid values: VND20,000, VND50,000, VND80,000, VND110,000
236 and VND130,000 (which were equivalent to US\$¹0.87/month, US\$2.18/month, US\$3.49/month,
237 US\$4.81/month, and US\$5.68/month). The closed-ended question format was chosen because of
238 its significant benefits over other forms. In fact, this type of question is much easier for respondents
239 to answer as they can focus on the question rather than the framing issues (Whitehead, 2006).
240 These bid values were based on additional discussions with local waste managers and public
241 officials from urban joint-stock companies as the local authority managing waste in every province
242 in Vietnam, and reflect the range of expected low and high-cost changes. Finally, respondents
243 were asked to provide some social demographic characteristics and their perceptions and benefits
244 of MSW programs. The commonly used model to estimate utility functions in CVM method is the
245 Logit model. This model is based on limited assumptions and is popular due to the simplicity of
246 estimations.

247 By contrast, the CE survey instrument included more detail about the proposed program
248 attributes and how changes to the environment might alter under different levels of those attributes.
249 When designing a CE survey both the key attributes and their relevant levels must be carefully

¹ US\$1 was equal to 22,890 Vietnamese Dong (VND) on June 30th, 2021

250 developed and tested (Pearce et al., 2006). The total number of attributes that can be taken into
1
2 251 account is proportional to the number of study observations (Bateman et al., 2002); the more data
3
4 252 gathered, the higher the number of attributes that can be included. However, to avoid complicated
5
6
7 253 choice sets for respondents the number of attributes should be limited (Hanley et al., 2002)—
8
9
10 254 typically to six attributes/levels at most inclusive of the payment option. Based on the discussions
11
12 255 and focus groups outlined above, we settled on four attributes for the study (with levels in
13
14 256 parentheses):

- 16 257 1. The percentage of solid waste recycled (0%, 5%, 10% and 15%) through the new program
- 18
19 258 2. The rate of CO₂ emission reductions (0%, 5%, 10% and 15%) achieved
- 20
21
22 259 3. The number of MSW separation classes on offer (unsorted, recycled + residual, or recycled
23
24 260 + organic + residual) through provided bins, and
- 25
26
27 261 4. The monthly fee for MSW management service (US\$0.86/month, US\$2.14/month,
28
29 262 US\$3.43/month, US\$4.71/month and US\$5.57/month).
- 30

31 263 Consistent with the literature, when investigating households' preferences for waste separation
32
33
34 264 these four attributes may differ according to specific local conditions (e.g. payment fee and the
35
36 265 number of MSW sorting classes) (Czajkowski et al., 2014). These attributes have been widely,
37
38
39 266 and successfully, used in different contexts. Therefore, their application for waste sorting in this
40
41 267 study is deemed appropriate. Naturally though, the attribute levels employed by our study differ
42
43
44 268 from those of others to accommodate local MRD household preferences for emission rate
45
46 269 reductions, levels of waste recycled, and sorted-waste classifications based on our pilot program
47
48
49 270 data (i.e. consistent with the government-provided bin classes).

50
51 271 Following the pre-testing phase, an orthogonal combination approach was used to create 25
52
53 272 alternative survey versions using the four attributes/levels detailed above. These options were then
54
55
56 273 incorporated into five final different survey versions (see Table 1 for an example choice set), and
57
58 274 each household was provided with one of those five survey versions. Each package corresponded
59
60
61
62
63
64
65

1 275 to a set of choices, and each question had three options. Option A and option B were hypothetical
2 276 solid waste management service quality improvement plans defined by the four criteria above,
3
4 277 with varying levels the attributes. Option C was the status quo option, where an individual saw no
5
6
7 278 need to increase the current MSW management service quality.
8







9 279 *4.2 Data collection*

10
11 280 Once finalized, the two survey instruments were implemented among a stratified sample of the
12
13
14 281 population. Following Jin et al. (2018) and Contu and Mourato (2020) we administered both
15
16 282 surveys to each household. This was done so that we could compare welfare estimates and
17
18
19 283 contribute to the ongoing examination of comparability between the two approaches in the
20
21
22 284 literature (see for example Lehtonen et al. 2003), and highlight better how each can inform
23
24 285 program or policy decision-making processes in developing contexts (Boxall et al. 1996). In order
25
26 286 to avoid bias when presented to respondents with CVM and CE on the same day, each scenario
27
28
29 287 was carefully explained to the respondents (Boxall et al. 1996).
30

31
32 288 The final data (380 observations) for our study was gathered by conducting direct interviews
33
34 289 with local residents in three of the biggest MRD cities: Can Tho (urban grade 1 – 146 responses),
35
36 290 Long Xuyen (urban grade 2 – 120 responses), and Ca Mau (urban grade 3 – 114 responses).
37
38
39 291 According to the Ministry of Natural Resources and Environment of Vietnam (2016), urbanization
40
41 292 is a critical reason for sharp recent solid waste increases; but where differing levels of urbanization
42
43
44 293 will drive different sources of waste and separation requirements. Thus, in urban settings within
45
46 294 developing countries, the use of stratified samples is common (see for example Chaudhry et al.
47
48 295 2007) and, for our purposes, useful as the results of this study may in future be applied to the rest
49
50
51 296 of the Mekong River Delta region.
52

53 297
54
55
56
57
58
59
60
61
62
63
64
65

298 **Table 1. Description of different version options in the CE questionnaire**

Categories	Option A	Option B	Option C	
Rate of MSW being recycled	Reduced 10% 	Reduce 5% 	Do not choose both A and B (Status quo)	
The rate of CO ₂ emissions is reduced	Reduced 15% 	Reduce 15% 		
Number of types of classified MSW	Recycling and remaining 	Non-classified 		
Fees for solid waste management service	US\$0.87/month	US\$2.18/month		
Please tick only 1 of the 3 options	<input type="checkbox"/>	<input type="checkbox"/>		<input type="checkbox"/>

299 **4.3 Data analysis**

300 The CVM survey instrument used in this study is based on random utility theory as detailed by
 301 Luce (1959) and McFadden (1973), which states that the indirect utility function as a vector of a
 302 households' use of a resource (V) follows the form:

$$V(p, q_i, M, \epsilon) \quad (1)$$

303 where p is the price vector, q is the number of goods, M is income, and ϵ is a random error. For
 304 simplicity, we can remove the price vector from the indirect utility function and assume that any
 305 change in an environmental good will be at the expense of a households' capacity to perform MSW
 306 source separation. Accordingly, when undertaking the classifying and separating their MSW, the
 307 utility of a household will be:

309
$$V(q_0, M + t_k, \varepsilon) \geq V(q_0, M, \varepsilon) \quad (2)$$

1
2 310 In the CVM scenario, the bid or cost of the program arose from our earlier discussions with
3
4 311 experts and residents. The probability that a household chooses to answer "Yes" with the *bid* t_k
5
6
7 312 is:

8
9 313
$$Pr[Yes] = Pr[V(q_0, M + t_k, \varepsilon_1) \geq V(q_0, M, \varepsilon_0)] \quad (3)$$

10
11 314 If we assume that the utility function is linear: $v(q_i, M) + \varepsilon_i$, (4)

12
13
14 315 then it is possible to write the probability formula (3) for the option "Yes" as:

15
16 316
$$Pr[Yes] = Pr[v(q_0, M + t_k) - v(q_0, M) + \varepsilon_1 - \varepsilon_0 \geq 0]. \quad (5)$$

17
18
19 317 The household will select "Yes" when the total usefulness changes, $\Delta U = v(q_0, M + t_k) - v(q_0, M)$,
20
21 318 and the difference in error, $\eta = \varepsilon_1 - \varepsilon_0$, is greater than 0. The subsequent probability is:

22
23 319
$$Pr[Yes] = Pr[\eta \geq -\Delta U]. \quad (6)$$

24
25
26 320 Based on the theory of probability, we have:

27
28 321
$$Pr[Yes] = Pr[\eta \geq -\Delta U] = 1 - F_\eta(-\Delta U), \quad (7)$$

29
30 322 where F_η is the cumulative density function (CDF) of η . If $F(x)$ has a symmetrical distribution then
31
32
33 323 $F(x) = 1 - F(-x)$. If we assume that η has a symmetrical distribution we can write:

34
35 324
$$Pr[Yes] = F_\eta(\Delta U). \quad (8)$$

36
37
38 325 Based on probability theory, the maximum likelihood estimation approach was employed to
39
40
41 326 determine the cumulative density function (CDF) and then the values of the corresponding
42
43
44 327 coefficients. A parametric method is employed to estimate willingness-to-pay mean and median
45
46 328 based on the coefficient of bid and other variable coefficients related to attitude and other
47
48
49 329 households' socio-economic characteristics. This research used the Logit model, which is one of
50
51 330 the approaches commonly used to estimate the cumulative density function when the random error
52
53 331 has a normal distribution, to estimate the coefficients of these variables. The Logit model is
54
55
56 332 presented as follows:

57
58 333
$$P_i = F(x_i'\beta) = \frac{e^{x_i'\beta}}{1 + e^{x_i'\beta}}, \quad (9)$$

334 where $x_i\beta = \beta_0 + \beta_1x_1 + \beta_2x_2 + \beta_3d_1 + \beta_4x_3 + \beta_5d_2 + \beta_6d_3 + \beta_7d_4 + \beta_8d_5$. (10)

335 The dependent variable (Y) is the probability of agreeing to pay for the increased MSW
336 management service fee. This variable holds two values: Y = 1 if the respondent agrees to pay,
337 and Y = 0 if the respondent does not agree to pay. The independent variables include the bid (X1),
338 which is the monthly fee per household for increased MSW management services. These rates are
339 suggested based on rates for unsubsidized MSW services from the government and local experts.
340 The variable age (X2) is the respondent's age (in years). The variable male (D_1) is a dummy
341 variable that holds two values: $D_1 = 1$ if the respondent's gender is male, and $D_1 = 0$ if the
342 respondent's gender is female. Edu (X3) is the respondent's years in school (in years). It is expected
343 that respondents with higher educational attainment tend to understand the benefits of increased
344 MSW management services and the environmental harm from solid waste. Therefore, they tend to
345 pay higher for any program changes. Income D_2 (dummy variable) holds two values: $D_2 = 1$, if the
346 respondent's monthly income is US\$385.44 or more (this is the level of a deduction for taxpayers
347 based on Law No. 26/2012/QH13 issued by the National Assembly of the Socialist Republic of
348 Vietnam on November 22, 2012), and $D_2 = 0$ when lower than US\$385.44. In this case, the demand
349 theory for environmental goods assumes that the higher the income, the better the demand for
350 environmental quality (Lewis and Tietenberg, 2019). The non-classified variable (D_3 - dummy
351 variable) takes two values: $D_3 = 1$, if the respondent does not recycle the MSW by classifying it,
352 and $D_3=0$ if otherwise. Finally, we also classify by type of urban setting, which differs with respect
353 to population size and density, rates of non-agricultural labour, and architectural or infrastructure
354 amenities (where type 1 > type 2 > type 3). The variable Urbantype2 (D_4 - dummy variable) again
355 takes two values: $D_4 = 1$, if the respondent is in urban city type 2, and $D_4 = 0$, if otherwise. Finally,
356 Urbantype3 (D_5 - dummy variable) has two values: $D_5 = 1$, if the respondent is in urban city type
357 3, and $D_5 = 0$ if otherwise.

358 The CE approach is based on the multi-criteria utility theory of Lancaster (1966) combined
 359 with the random utility theory of Thurstone (1927). The random utility theory states that an
 360 individual consumer's utility consists of observable and unobservable parts. The observable
 361 (measurable) portion of an individual's utility is based on their evaluation of the product attributes.
 362 The unobservable portion is random and depends on the individual's preferences. The utility
 363 function of an individual i when consuming product j is:

$$U_{ij} = V_{ij} + e_{ij} = V(Z_{ij}, S_i) + e(Z_{ij}, S_i) \quad (11)$$

364 Where V is the observable part. V_{ij} is a vector of the degree of the attributes Z of product j together
 365 with the economic, social and attitudinal characteristics (S) of respondent i , e is the unobservable
 366 part. Faced with a choice set consisting of many different products with different attributes,
 367 consumers will choose the product that gives them the maximum utility ($\max U$). The probability
 368 that individual i chooses product j over any other product m corresponds to the likelihood that U_j
 369 $> U_m$. Specifically, the probability of choosing product j of individual i (P_{ij}) will be:

$$P(i) = P(U_{ij} > U_{im}) = P(V_{ij} + e_{ij} > V_{im} + e_{im}); \forall m \in C \quad (12)$$

370 Assuming that the random component e_{ij} follows a homogeneous, independent and identical
 371 distribution (IID), and follows a Gumbell or Weibull distribution, the probability that **alternative**
 372 **j is selected is estimated using Multinomial Logit (MNL)** model as follows:

$$P(y_i = j | C) = \frac{\exp(Z'_{ij}\beta + S'_i\delta)}{\sum_{m \in C} \exp(Z'_{im}\beta + S'_i\delta)} \quad (13)$$

373 This study employed a Hausman and McFadden (1984) test to check whether the IIA property is
 374 violated. If the IIA property is violated, the random parameter logit (RPL) is then applied. The
 375 random utility function in the RPL model is as follows:

$$U_{ij} = V_{ij} + e_{ij} = Z'_{ij}\beta + S'_i\delta + Z'_{ij}\phi_i + e_{ij} \quad (14)$$

381 Where v is the deterministic component of the latent utility and e is the error component stochastic
 1
 2
 3 382 term, $\beta + \varphi_i$ is the population mean, and $\varphi_i \sim N(0, \mu_i)$ is the stochastic deviation that represents
 4
 5 383 the individual's preferences relative to the average preferences in the population. Considering
 6
 7 384 unconditional heterogeneity in preferences across respondents and conditional on the
 8
 9
 10
 11 385 unobservable φ_i (Birol et al., 2006), Eq. (13) now changes as follows:

$$386 \quad P(y_i = j | C)\varphi_i = \frac{\exp(Z'_{ij}\beta + S'_i\delta + Z'_{ij}\varphi_i)}{\sum_{m \in C} \exp(Z'_{im}\beta + S'_i\delta + Z'_{ij}\varphi_i)} \quad (15)$$

17
 18 387 The stochastic portion of the utility in this model may be correlated among alternatives and across
 19
 20
 21 388 the sequence of choices through the effect of φ_i because of an unrestricted IIA assumption. The
 22
 23 389 requirement of treating preference parameters as random variables is the estimation of the
 24
 25
 26 390 simulated maximum likelihood. Procedurally, the maximum likelihood algorithm searches for a
 27
 28 391 solution by simulating n draws from distributions with given means and standard deviations. The
 29
 30
 31 392 probabilities are estimated by integrating the joint simulated distribution.

32
 33 393 The linear equation of utility for the choice of the j^{th} product is written as:

$$394 \quad V_{ij} = ASC + \beta_1 Z_1 + \beta_2 Z_2 + \beta_3 Z_3 + \dots + \beta_k Z_k \quad (16)$$

35
 36
 37
 38 395 where k is the order of the product attributes. The coefficient β can be negative or positive,
 39
 40
 41 396 different for each product attribute, and is "valued" according to the subjective preferences of each
 42
 43 397 individual. The coefficient β will vary between groups of individuals in a population but be the
 44
 45
 46 398 same for individuals in the same group. Although there are many ways to remove the mismatch
 47
 48 399 from the choices and improve the model fit, this research uses the MNL model to reduce the error
 49
 50
 51 400 and give the most accurate results.

52
 53 401 Based on formula (16), the utility function of choice A, B, and C in each set of questions
 54
 55 402 can be presented as follows:

$$58 \quad 403 \quad \text{Option A} \quad V_1 = ASC + \beta_1 fee + \beta_2 waste + \beta_3 co2 + \beta_4 sep2 + \beta_5 sep3$$

404 Option B $V_2 = ASC + \beta_1 fee + \beta_2 waste + \beta_3 co2 + \beta_4 sep2 + \beta_5 sep3$

405 Option C $V_3 = \beta_1 fee + \beta_2 waste + \beta_3 co2 + \beta_4 sep2 + \beta_5 sep3$

406 where V_j is the utility function associated with choice j , and the ASC is a constant for each choice
 407 in the model. The marginal willingness to pay ($MWTP$) for improved properties via the proposed
 408 MSW program is estimated by the marginal rate of substitution between the non-monetary
 409 attribute parameter $\beta_{non-monetary\ attribute}$ and the monetary attribute factor $\beta_{monetary\ attribute}$ as
 410 follows:

$$411 \quad MWTP = - \frac{\beta_{non-monetary\ attribute}}{\beta_{monetary\ attribute}}$$

412 The coefficients $\beta_{non-monetary\ attribute}$ and $\beta_{monetary\ attribute}$ are estimated from the MNL
 413 model. The variables included in the MNL model are the attributes of the MSW management
 414 service quality improvement program. These variables include the variable fee , variable $waste$,
 415 $co2$, and two separation variables $sep2$ and $sep3$. The variable fee represents the charging attribute
 416 for increased MSW management service stated as the monthly payment the household will incur
 417 for the MSW service (US\$/month). The variable $waste$ (%) represents the percentage of recycled
 418 solid waste. The variable $co2$ is a variable representing an attribute of reduced CO₂ emissions (%)
 419 from changes to waste management, while the variables $sep2$ and $sep3$ represent numeric attributes
 420 of the possible types that MSW may need to be separated into (recycled, organic, residual). The
 421 variable $sep2$ is a dummy variable that takes two values: $sep2 = 1$, if the number of MSW types is
 422 classified into 2 types, recycled and remaining, and $sep2 = 0$ if the number of MSW types is not
 423 classified into any type (unclassified). The variable $sep3$ is another dummy variable with two
 424 values: $sep3 = 1$ if the number of MSW types is classified into three categories (recycled, organic,
 425 and residual), or $sep3 = 0$ if the MSW is not classified into those three categories.

426 *Welfare analysis*

427 In order to compare welfare measurements from each technique, the CE is restricted to evaluate
 428 the welfare impact of the same improvement presented in CVM. It means our CVM examined

429 only the MSW management program change, while the CE model estimates any welfare changes
430 relative to different attribute levels. For the CE the change of a suggested MSW management
431 program was valued using the following function (for further information see Boxall et al. 1996;
432 Morrison et. al, 1999; Jin et. al., 2018):

$$WTP = -\frac{1}{\beta_M}(V^0 - V^1),$$

434 where β_M is the coefficient of monetary attribute and is interpreted as the marginal utility income.
435 V^0 and V^1 represent the indirect observable utility before and after the change under consideration.

436 Data analysis in this paper was conducted using STATA version 14 and NLOGIT 5 statistical
437 software (package for choice modeling estimation).

438 5. Results and discussion

439 5.1 Socio-economic characteristics of the respondents

440 Thirty percent of the respondents were male, and 70% were female, with the average age of
441 respondents being 49.6 years old. About 1.32% of respondents said they did not go to school. Of
442 those that had attended school 23.95% had attended elementary school, 33.68% went to secondary
443 school, 30.79% went to high school, and 10.26% went on to higher studies. With respect to
444 employment the two biggest categories, traders and homemakers, represented 33.16% and 26.84%
445 of the sample respectively. Other employment categories included retirees (10.25%), government
446 roles (6.32%), and working for private enterprise (3.42%). The majority of those who responded
447 had an average monthly income of less than US\$385.44, accounting for 90.79% of those surveyed.

448 These socio-economic characteristics of respondents are consistent with population-level data
449 from the Vietnam Household Living Standard Survey (VHLSS) conducted by Vietnam General
450 Statistic Office (GSO) every two years. Therefore, the representativeness of the data in this survey
451 was confirmed.

452 *Attitudes of the respondents towards the situation of MSW*

453 Because the research was conducted in metropolitan areas, most participants received solid waste
 1
 2 454 collection services (98.42%). The average quantity of MSW emitted per residence is 1.84 kg per
 3
 4 455 day, which is collected by city sanitation workers. Although localities in the MRD have not yet
 5
 6
 7 456 implemented the program of classifying MSW at source our survey results showed that 50% of
 8
 9
 10 457 respondents were aware of the process behind at-source separation MSW programs (according to
 11
 12 458 the instructions of the Vietnam Government's Decree 38/2015/ND-CP on waste and scrap
 13
 14 459 management), and 67.11% have implemented it on the basis of being able to sell some waste
 15
 16 460 products as a small source of revenue. By contrast, 32.89% of households have no experience with
 17
 18
 19 461 selling waste goods to third parties with many complaining that the amount sold does not provide
 20
 21
 22 462 sufficient income to alter their behaviour. Finally, there were some respondents (2%) who stated
 23
 24 463 they had little concerns for or about waste classifying behavior.

26 464 *5.2 Estimation results from CVM and CE for the improved management program*

29 465 *Estimation results - CVM approach*

31 466 Our results found that 66.31% of respondents were willing to pay for improved MSW management
 32
 33 467 services which is a relatively high acceptance rate. Across the bid levels, 94.87% of respondents
 34
 35
 36 468 agreed to pay the monthly cost of US\$0.87; just 5.12% refused to pay. When the monthly charge
 37
 38
 39 469 was increased to \$2.18, 77.63% still agreed to pay. At the price of US\$3.49/month, the number of
 40
 41 470 those who decided to pay fell to 65.27%. When the charge was raised to US\$4.81/month, the
 42
 43
 44 471 acceptance rate dropped to 53.84%. Only 39.47% of respondents agreed to pay the maximum
 45
 46 472 monthly cost of US\$5.68. These outcomes are compatible with demand curve economic theory
 47
 48
 49 473 and as we would expect—but encouraging given the positive levels of participation among
 50
 51 474 residents at the higher payment levels.

53 475 **Table 2. The proportion of willingness to pay for improved MSW management services**

Bid	Willing to pay	Unwilling to pay
-----	----------------	------------------

(US\$/month)	Total of observations	No. of respondents	Proportion (%)	No. of respondents	Proportion (%)
0.87	78	74	94.87	4	5.12
2.18	76	59	77.63	17	22.36
3.49	72	47	65.27	25	34.72
4.81	78	42	53.84	36	46.15
5.68	76	30	39.47	46	60.52
Total	380	252	66.31	128	33.68

Source: Survey data, 2020

Table 3 summarizes the findings of the Logit model used to estimate willingness to pay based on the respondents' socio-economic characteristics, where the model's prediction percentage is 76.58% which is both reasonable and acceptable. We next grouped households into three categories: Group 1 comprised households with previous waste sorting experience, Group 2 households did not have any prior experience in waste separation, and Group 3 combined both into a single collective. The main purpose of this exercise was to determine if significant differences in WTP values exist between groups. Interestingly, Group 2 respondents reported higher WTP for waste separation despite no prior engagement with such activity (Table 4), which may reflect a distaste for such activity generally. Overall, the determinants of the respective WTP are consistent across our three models, but there are statistically significant differences between experienced waste-sorting households and those that have never undertaken this activity. The regression findings also demonstrate that the parameter of the Bid variable (significant at the 1% level) has a negative sign, indicating that the higher the charge for the MSW management service, the lower the chance of agreeing to pay. This result is consistent with economic theory and people's preference for recycling programs, depending on how much it would cost their households to participate in the programs (Kipperberg and Larson, 2012). Furthermore, the data reveals that the

493 older the respondent the less likely they are WTP for increased services, possibly due to limited or
 494 fixed income constraints. This may be confirmed by the fact that those with higher incomes are
 495 more likely to pay for the proposed changes. These results conform with previous research by
 496 Rahji and Oloruntoba (2009), Pek and Othman (2010), and Altaf and Deshazo (1996), in which,
 497 age, income, family size, and employment have all been identified as socio-economic predictors
 498 of household waste handling behavior. Our results also reveal important factors which previous
 499 studies have not mentioned, where living conditions and waste types would give people different
 500 incentives to recycle. This result is somewhat consistent with Heller and Vatn's (2017) results, in
 501 which the authors argue that one of the motivations for recycling is an individual's concern for the
 502 environment.

503 **Table 3. Logit regression results on factors affecting the willingness to pay for the**
 504 **improvement of the MSW management program**

Variables	Model 1 - Group 1		Model 2 – Group 2		Model 3 - Combined group	
	Coef.	z value	Coef.	z value	Coef.	z value
Constant	1.67568***	3.15	2.72104***	2.76	1.75397***	3.81
Bid	-0.000017***	-6.85	-0.000016***	-4.13	-0.000017***	-8.03
Age	-0.01263*	-1.67	-0.02575*	-2.10	-0.01549**	-2.45
Male	-0.10289 ^{ns}	-0.50	0.57605*	1.75	0.09211 ^{ns}	0.54
Edu	0.02973 ^{ns}	1.22	-0.01061 ^{ns}	-0,23	0.02513 ^{ns}	1.18
Income	0.95496**	2.17	1.17246**	2.27	1.02106***	3.12
Non-classified					0.31167*	1.85
Urbantype2	0.61853***	2.75	0.56527 ^{ns}	1.48	0.63748***	3.04
Urbantype3	0.96389***	4.00	0.90777**	2.52	0.94405***	4.79
Log-likelihood	-128.48043		-54.00449		-184.55692	
LR chi ² (8)	76.51		42.08		116.46	

Prob > chi ²	0.0000	0.0000	0.0000
Pseudo R ²	0.2294	0.2804	0.2398
Observation	255	125	380

505 *Source: Survey data, 2020*

506 *Note: *, **, and *** are statistically significant at 10%, 5%, and 1%, respectively, and ^{ns} are not statistically significant*

507 *Group 1: Households with previously sorted garbage, otherwise belong to Group 2. Sorted garbage indicates*

508 *activities carried out by households, in which several recycling materials such as paper, metal, and plastic bottles*

509 *are separated and then sell to informal individual waste collectors.*

510 Respondents that did not perform MSW separation stated that this was because they did not
511 have time to do so, or because sorted MSW required a lot of storage space. This set of respondents
512 was also more likely to pay higher rates for change than others. Further, respondents in Type 2
513 and 3 urban regions were more likely to agree to pay than respondents in Type 1 urban regions.
514 This may be explained by the fact that existing MSW management systems in Type 1 cities are
515 more thorough than those in other urban settings and, as a result, respondents in Type 2 and 3
516 cities have higher expectations for increased MSW management services that are expected to
517 contribute to environmental improvement.

518 The mean WTP elicited from the CVM and CE surveys are presented in Table 4. The
519 parameter findings from the Logit model suggest that the average WTP value for increased MSW
520 management via the CVM is US\$4.81/month/household, ranging from US\$4.49/month/household
521 to US\$5.47/month/household (at 1% significance level). Meanwhile, the mean WTP from the CE
522 survey estimates is about US\$4.61; or a little smaller than the CVM average. As stated above, the
523 mean WTP for Group 1 respondents is smaller than that of other groups.

524 **Table 4. Willingness to pay mean value estimated from CVM and CE**

		MWTP	Lower Bound	Upper Bound	ASL
Group 1	CVM	4.62	4.16	5.25	0.0000

	CE	4.54 (1.80)	4.04	5.04	
Group 2	CVM	5.58	4.75	7.38	0.0002
	CE	4.84 (1.83)	4.33	5.35	
Combined group	CVM	4.81	4.49	5.47	0.0000
	CE	4.61 (1.81)	4.11	5.11	

Note: ASL is the significance level for hypothesis tests: $H_0: WTP \leq 0$, $H_1: WTP > 0$, number in parentheses are std. err.

5.3 Estimation results - CE approach

Table 5 reveals the results of the CE approach with MNL estimation. The results indicate that all the attributes are statistically significant except Sep2 and Sep3. The results show that the attribute variable fee, waste type and CO₂ influence people's willingness to pay for the program to improve the quality of MSW management services. The coefficient of the fee variable has a negative value, showing a negative correlation to the people's decision to pay. This is consistent with the law of demand and price effects. In contrast to the fee variable, the coefficients of the waste and CO₂ variables have a positive value, increasing the decision to pay for an improvement in MSW management services, and the probability of people participating in the program. However, the final fee value for the program should be calculated at a reasonable level because this is a factor that reduces the probability of people participating.

Table 5. Estimated results of the MNL model

Variable name	Coeff.	z value
ASC	2,70137***	6,94
Fee	-0,000023337***	-16,92
Waste	1,53366*	1,82
CO ₂	3,09356***	3,61
Sep2	0,11476 ^{ns}	1,63
Sep3	0,05827 ^{ns}	0,70
Log-likelihood	-1.161,55238	
Observation	1,965	

Source: Survey data, 2020

540 The test results of IIA property are summarised in Table 6. The values of the test statistics
 541 are significantly greater than the critical value of the Chi-square distribution, which is 18.475 at
 542 the 1% significance level with seven degrees of freedom. Therefore, the null hypothesis of the IIA
 543 restriction is rejected. However, the constant variance assumption would be violated if the
 544 alternatives Rice A and Rice B were dropped from the choice sets; indicating that using the CL
 545 model approach to analyze this data may not be appropriate in terms of the IIA assumption. A less-
 546 restrictive specification of the choice model was thus considered to obtain unbiased and better
 547 results via an RPL model, with the results shown in Table 7.

548 **Table 6. IIA Test results**

Alternative dropped	χ^2	Degree of F	Probability
Alternative A	348,0945	6	0.0000
Alternative B	3.918,3110	6	0,0000

549 *Source: Survey data, 2020*

550 The results of the RPL regression model reveal the attribute determinants of any WTP decision in
 551 support of improved MSW management services (Table 7). Across all *three models* the attribute
 552 variables charge, waste, *co2*, *sep2*, and *sep3* impacted people's willingness to pay for a program
 553 to increase. The coefficient of the fee variable has a negative sign, indicating that the charge has a
 554 negative link with the respondents' payment decision. The coefficients for waste, *co2*, *sep2*, and
 555 *sep3* on the other hand have a positive sign, indicating a favorable link to the choice to pay more
 556 to increase MSW management services. As a result, when the MSW management program is
 557 improved by gradually raising the proportion of recycled MSW, reducing the rate of CO₂ emitted
 558 from MSW treatment, and categorizing additional categories of MSW, people are more inclined
 559 to participate in the program.

560 **Table 7. Estimated results of RPL model on the determinant of households' willingness to**
 561 **pay for source-separation of solid waste**

Model 1 - Group 1	Model 2 – Group 2	Model 3 - Combined group
-------------------	-------------------	--------------------------

Variable name	Coeff.	z value	Coeff.	z value	Coeff.	z value
Fee	-0.000088***	-9.09	-0.000081***	-5.65	-0.000089***	-10.89
Waste	11.07539***	6.05	8.69775***	3.20	10.38837***	6.61
CO ₂	17.07812***	6.29	22.96321***	4.45	19.61559***	7.74
Sep2	0.40276***	2.70	0.55241**	2.42	0.47944***	3.73
Sep3	0.20030 ^{ns}	0.95	1.07653***	3.40	0.50683***	2.86
Log-likelihood	-698.86745		-325.51282		-1,029.320	
Observations	4,065		1,812		5,877	
Chi ²	222.36		119.38		348.39	
Prob > Chi ²	0.0000		0.0000		0.0000	

562

563 However, the payment charge for program improvements should be calculated at an appropriate
564 level because this is a factor that may reduce the likelihood of resident participation. Further, our
565 results provide evidence of a strong positive relationship between *sep3* and WTP by Group 2
566 members as well as those in the combined Group 3. This effect is not statistically significant for
567 Group 1. Therefore, at source separation of MSW into the three categories (recycled, organic and
568 residual) is not significantly correlated with an increased WTP. The MWTP values for each of the
569 program's attribute characteristics were generated based on the RPL regression findings (Table 8).

570 Table 8 illustrates that individuals are interested in the attributes focused on in our research.
571 Interestingly, Group 2 has the highest mean WTP for three of the attributes (*co2*, *sep2*, *sep3*) and
572 the lowest MWTP for the *waste* attribute. This result conforms with the CVM findings. It is
573 possible that this result is reflective of the perceptions and characteristics of households on MSW
574 separation, as explained above, where households with lower education and income may not

575 expect many environmental benefits (or attributes) from this behavioural change. However, chief
 576 among these attributes is a reduction in CO₂ emissions, where the marginal willingness to pay for
 577 the attribute (US\$9.61/month) is nearly twice as high as the marginal willingness to pay for the
 578 attribute of increasing recycling rate (US\$5.09/month) and forty times the marginal willingness to
 579 pay for separate types of solid waste. This result is consistent with previous studies (e.g., Sakata,
 580 2007), which found that individuals in Japan are prepared to pay a premium to minimize dioxin
 581 emissions. This result also suggests that the willingness to pay for the program's features is more
 582 than the present charge of around US\$1.13/month/household or US\$3.40/year/person, indicating
 583 a favorable indicator for the enhancement of the quality of MSW management services (World
 584 Bank, 2018). Environmental attributes have been considered important factors that determine
 585 household willingness to contribute to sorting programs. Consistent with previous studies, this
 586 research indicates that concerns related to waste and CO₂ increase household willingness.
 587 Furthermore, Sep2 and Sep3 which are related to environmental concerns also motivate higher
 588 willingness (Heller and Vatn (2017). Once again though, the cost of the program associated with
 589 recycling will be critical, where the fee variable clearly shows a negative influence on WTP.

590 **Table 8. WTP for program attributes (US\$/month)**

Variables	MWTP		
	Group 1	Group 2	Combined group
Waste	5.48***	4.65***	5.09***
	(3.71 to 7.24)	(1.91 to 7.39)	(3.60 to 6.58)
CO ₂	8.45***	12.27***	9.61***
	(6.17 to 10.71)	(8.03 to 16.52)	(7.62 to 11.60)
Sep2	0.20***	0.30**	0.23***
	(0.06 to 0.34)	(0.06 to 0.53)	(0.11 to 0.36)

Sep3	0.10 ^{ns}	0.57 ^{***}	0.25 ^{***}
	(0.10 to 0.31)	(0.27 to 0.89)	(0.08 to 0.42)

591

592 More importantly, as the current fee structure is insufficient to cover the existing collection,

593 transportation, and treatment of solid waste where the total cost is US\$1.68/month/person (World

594 Bank, 2018) the opportunity to reduce the financial burden on the state budget is significant. This

595 is important for a country such as Vietnam where economic development is essential but budget

596 constraints are typically high. According to the Vietnam Ministry of Finance (2015), state budget

597 expenditures on MSW management have more than quadrupled from US\$266,809.42 in 2010 to

598 US\$488,222.70 in 2015. The capacity to address that shortfall, and aid policymakers to implement

599 Prime Ministerial Directive 33 (2020) as a result of this research, offers an important contribution

600 toward expanding waste management and reuse capacity in Vietnam.

601 **6. Conclusions and policy implications**

602 Finding from this paper should be interesting and useful to policy-makers, especially in developing

603 countries that are struggling to implement and maintain effective MSW programs. We find

604 positive rates of WTP for increased MSW management services among those that have previously

605 engaged with waste separation activity and those that have not; where interestingly those that have

606 not undertaken the behaviour previously may be willing to pay higher rates for the service. This

607 may indicate some reluctance to engage in separation activity for revenue purposes (where such

608 income generation is possible), but still manifest as a willingness to participate in waste separation

609 for collection purposes. This specific finding may have to be tested further through focus sessions

610 and additional interviews with members from that group. However, we do find that separation

611 services offered under any increased MSW program are correlated with an increased willingness

612 by residents to pay for that service.

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
613 The WTP estimates are broadly consistent across the two survey approaches and show no
614 statistically significant difference. This finding is consistent with the limited number of previous
615 studies that have compared welfare estimates between these two approaches. An additional driver
616 of higher WTP appears to be any reduction to CO₂ emissions provided by environmental changes
617 under a changed program. Environmental concerns have grown in relevance for Vietnamese
618 residents in recent years, likely driving some of the findings reported in this study. Responses
619 collected via the CVM survey suggested an average willingness to pay of US\$4.81/month, while
620 the CE survey WTP estimate was US\$4.61/month—again consistent with other studies that show
621 CVM estimates can often be higher than those of CE. However, the difference here is negligible
622 as shown and offers a clear price range for policy-makers to consider adopting; and one that is
623 well above the current US\$0.86/month revenue for existing services, and also higher than the
624 US\$1.68/month real costs of Vietnamese MSW operations.

29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65
625 When implementing the proposed program changes it will be critical for management
626 agencies to strengthen their communication programs in order to raise awareness of environmental
627 benefits that can stem from improved MSW and the source-separation of household waste. These
628 awareness-raising activities should initially focus on households with higher incomes that already
629 actively participate in environmental-related activities in their areas. However, at the margin,
630 further useful policy advice has been provided by our study from discrepancies between the CVM
631 and CE welfare estimate approaches. While the CVM survey determined an economic worth of
632 changes to the existing solid waste management program, by contrast, the CE survey enabled us
633 (and Vietnam's policy-makers) to estimate the economic worth of key MSW management
634 program attributes. Respondents, in this case, have reported a willingness to engage in separation
635 behaviour, and pay for the MSW services required to collect, process and reuse that waste. This is
636 an important finding, as recycling habits are frequently seen as a component of everyday activities
637 that are significantly influenced by patterns; where further government efforts to incentivize those

638 who would prefer the status quo (~33% in this case) may assist in motivating change and full
1
2 639 adoption of improved MSW practices. This could involve rewards for those that do the right thing,
3
4 640 or neighborhood competitions related to waste separation activity. In any case, our research
5
6
7 641 provides useful insight for governments in developing countries to manage the growing problem
8
9
10 642 of waste management and recycling needs into the future.

11
12 643 **Acknowledgment:** This study is funded in part by the Can Tho University Improvement Project
13
14 644 VN14-P6, supported by a Japanese ODA loan.

16 17 645 7. References

- 18
19 646 Aadland, D., & Caplan, A. J., 2006. Curbside recycling: Waste resource or waste of
20
21 647 resources?. *Journal of Policy Analysis and Management: The Journal of the Association for*
22
23
24 648 *Public Policy Analysis and Management.* 25(4), 855-874.
- 25
26 649 Adamowicz, W., Boxall, P., Williams, M., & Louviere, J., 1998. Stated preference approaches for
27
28
29 650 measuring passive use values: choice experiments and contingent valuation. *American*
30
31 651 *Journal of Agricultural Economics.* 80(1), 64-75.
- 32
33
34 652 Adeoti, A., & Obidi, B., 2010. Poverty and preference for improved solid waste management
35
36 653 attributes in Delta-State, Nigeria. *Journal of Rural Economics and Development.* 19(1623-
37
38
39 654 2016-134902), 15-33.
- 40
41 655 Altaf, M. A., & Deshazo, J. R., 1996. Household demand for improved solid waste management:
42
43 656 A case study of Gujranwala, Pakistan. *World Development.* 24(5), 857-868.
- 44
45
46 657 Anaman, K.A., Jair, R.M., 2000. Contingent valuation of solid waste collection services for rural
47
48
49 658 households in Brunei Darussalam. *The Singapore Economic Review.* 45, 223-240.
- 50
51 659 Arrow, K.J., 2001. Foreword, in: Bateman, I., Willis, K. (Eds.), *Valuing environmental*
52
53 660 *preferences: theory and practice of the contingent valuation method in the US, EU, and*
54
55
56 661 *developing countries.* Oxford University Press on Demand, New York, pp. 1-14.
- 57
58
59
60
61
62
63
64
65

- 662 Baldassare, M., & Katz, C., 1992. The personal threat of environmental problems as predictor of
1
2 663 environmental practices. *Environment and behavior*, 24(5), 602-616.
3
4
5 664 Bateman, I. J., Carson, R. T., Day, B., Hanemann, M., Hanley, N., Hett, T., & Sugden, R., 2002.
6
7 665 Economic valuation with stated preference techniques: a manual. Economic valuation with
8
9
10 666 stated preference techniques: a manual.
11
12 667 Boxall, P. C., Adamowicz, W. L., Swait, J., Williams, M., & Louviere, J., 1996. A comparison of
13
14 668 stated preference methods for environmental valuation. *Ecological economics*. 18(3), 243-
15
16
17 669 253.
18
19 670 Chaudhry, P., Singh, B. and Tewari, V.P., 2007. Non-market economic valuation in developing
20
21 671 countries: Role of participant observation method in CVM analysis. *Journal of Forest*
22
23
24 672 *Economics*, 13(4), pp.259-275.
25
26
27 673 Contu, D., Mourato, S., 2020. Complementing choice experiment with contingent valuation data:
28
29 674 Individual preferences and views towards IV generation nuclear energy in the UK. *Energy*
30
31 675 *Policy* 136, 111032.
32
33
34 676 Czajkowski, M., Kądziała, T., & Hanley, N., 2014. We want to sort! Assessing households'
35
36 677 preferences for sorting waste. *Resource and energy economics*. 36(1), 290-306.
37
38
39 678 Czajkowski, M., Hanley, N., & Nyborg, K., 2014. Social norms, morals and self-interest as
40
41 679 determinants of pro-environment behaviours (No. 18/2014). *Memorandum*.
42
43
44 680 Damigos, D., Kaliampakos, D., & Menegaki, M., 2016. How much are people willing to pay for
45
46 681 efficient waste management schemes? A benefit transfer application. *Waste Management &*
47
48 682 *Research*. 34(4), 345-355.
49
50
51 683 Das, P., Mukherjee, S., & Sen, R., 2008. Improved bioavailability and biodegradation of a model
52
53 684 polyaromatic hydrocarbon by a biosurfactant producing bacterium of marine
54
55
56 685 origin. *Chemosphere*, 72(9), 1229-1234.
57
58
59
60
61
62
63
64
65

- 686 De Young, R., 1986. Some psychological aspects of recycling: the structure of conservation-
1
2 687 satisfactions. *Environment and behavior*, 18(4), 435-449.
3
4
5 688 Diafas, I., 2016. Estimating the Economic Value of forest ecosystem services using stated
6
7 689 preference methods: the case of Kakamega forest, Kenya (Doctoral dissertation,
8
9 690 Niedersächsische Staats-und Universitätsbibliothek Göttingen).
10
11
12 691 Diamond, P. A., & Hausman, J. A., 1994. Contingent valuation: is some number better than no
13
14 692 number?. *Journal of economic perspectives*. 8(4), 45-64.
15
16
17 693 Ferreira, S., & Marques, R. C., 2015. Contingent valuation method applied to waste
18
19 694 management. *Resources, Conservation and Recycling*. 99, 111-117.
20
21
22 695 Fukuda, K., Isdwiyani, R., Kawata, K., & Yoshida, Y., 2018. Measuring the impact of modern
23
24 696 waste collection and processing service attributes on residents' acceptance of waste
25
26 697 separation policy using a randomised conjoint field experiment in Yogyakarta Province,
27
28 698 Indonesia. *Waste Management & Research*. 36(9), 841-848.
29
30
31 699 Garrod, G., Willis, K., 1998. Estimating lost amenity due to landfill waste disposal. *Resources,*
32
33 700 *Conservation and Recycling*. 22, 83-95.
34
35
36 701 Ghanbari, F., Sharee, F.A., Monavari, M., Zaredar, N., 2012. A new method for environmental
37
38 702 site assessment of urban solid waste landfills. *Environmental Monitoring and Assessment*.
39
40 703 184, 1221-1230.
41
42
43 704 Gillespie, R., & Bennett, J., 2013. Willingness to pay for kerbside recycling in Brisbane,
44
45 705 Australia. *Journal of environmental planning and management*. 56(3), 362-377.
46
47
48 706 Government of Vietnam, 2007. Decree 59 on solid waste management 59/2007/ND-CP dated
49
50 707 April 09, 2007. Hanoi.
51
52
53 708 Government of Vietnam, 2020. Directive No. 33/CT-TTg of the Prime Minister: On strengthening
54
55 709 the management, reuse, recycling, treatment and reduction of plastic waste. Hanoi.
56
57
58
59
60
61
62
63
64
65

- 710 Hanley, N., MacMillan, D., Wright, R. E., Bullock, C., Simpson, I., Parsisson, D., & Crabtree, B.,
1
2 711 1998a. Contingent valuation versus choice experiments: estimating the benefits of
3
4 712 environmentally sensitive areas in Scotland. *Journal of agricultural economics*. 49(1), 1-15.
5
6
7 713 Hanley, N., Wright, R. E., & Adamowicz, V., 1998b. Using choice experiments to value the
8
9 714 environment. *Environmental and resource economics*. 11(3), 413-428.
10
11
12 715 Hanley, N., Wright, R. E., & Koop, G., 2002. Modelling recreation demand using choice
13
14 716 experiments: climbing in Scotland. *Environmental and resource Economics*. 22(3), 449-
15
16 717 466.
17
18
19 718 Jin, J., Wang, Z., & Ran, S., 2006. Comparison of contingent valuation and choice experiment in
20
21 719 solid waste management programs in Macao. *Ecological Economics*. 57(3), 430-441.
22
23
24 720 Jin, J., He, R., Wang, W., & Gong, H., 2018. Valuing cultivated land protection: A contingent
25
26 721 valuation and choice experiment study in China. *Land use policy*. 74, 214-219.
27
28
29 722 Jin, J., Wang, Z., & Ran, S., 2006. Estimating the public preferences for solid waste management
30
31 723 programmes using choice experiments in Macao. *Waste management & research*, 24(4),
32
33 724 301-309.
34
35
36 725 Johnston, R., Boyle, K.J., Adamowicz, W., Bennett, J., Brouwer, R., Cameron, T.A., Hanemann,
37
38 726 W.M., Hanley, N., Ryan, M., Scarpa, R., Tourangeau, R., Vossler, C.A., 2017.
39
40 727 Contemporary guidance for stated preference studies. *Journal of the Association of*
41
42 728 *Environmental and Resource Economists* 4, 319-405.
43
44
45 729 Karousakis, K., & Birol, E., 2008. Investigating household preferences for kerbside recycling
46
47 730 services in London: A choice experiment approach. *Journal of environmental*
48
49 731 *management*, 88(4), 1099-1108.
50
51
52 732 Kipperberg, G., & Larson, D. M., 2012. Heterogeneous preferences for community recycling
53
54 733 programs. *Environmental and Resource Economics*, 53(4), 577-604.
55
56
57
58
59
60
61
62
63
64
65

- 734 Ko, S., Kim, W., Shin, S.-C., Shin, J., 2020. The economic value of sustainable recycling and
1
2 735 waste management policies: The case of a waste management crisis in South Korea. *Waste*
3
4 736 *Management*, 104, 220-227.
- 737 Lancaster, K. J., 1966. A new approach to consumer theory. *Journal of political economy*. 74(2),
8
9 738 132-157.
- 10
11
12 739 Lee, M., Choi, H., & Koo, Y., 2017. Inconvenience cost of waste disposal behavior in South
13
14 740 Korea. *Ecological Economics*, 140, 58-65.
- 15
16
17 741 Lehtonen, E., Kuuluvainen, J., Pouta, E., Rekola, M., Li, C.-Z., 2003. Non-market benefits of
18
19 742 forest conservation in southern Finland. *Environmental Science & Policy* 6, 195-204.
- 20
21
22 743 Lewis, L., & Tietenberg, T., 2019. *Environmental economics and policy*. Routledge.
- 23
24 744 Luce, R. D., 1959. *Individual Choice Behavior a Theoretical Analysis*, John Wiley and sons
- 25
26 745 Maimoun, M. A., Reinhart, D. R., & Madani, K., 2016. An environmental-economic assessment
27
28 746 of residential curbside collection programs in Central Florida. *Waste management*. 54, 27-
29
30 747 38.
- 31
32
33
34 748 Massarutto, A., Marangon, F., Troiano, S., & Favot, M., 2019. Moral duty, warm glow or self-
35
36 749 interest? A choice experiment study on motivations for domestic garbage sorting in
37
38 750 Italy. *Journal of cleaner production*, 208, 916-923.
- 39
40
41 751 McFadden, D. "Conditional Logit Analysis of Qualitative Choice Behavior" in Zarembka, P
42
43 752 (ed.), *Frontiers in Econometrics* (New York: Academic Press, 1974).
- 44
45
46 753 Miafodzyeva, S., & Brandt, N., 2013. Recycling behaviour among householders: Synthesizing
47
48 754 determinants via a meta-analysis. *Waste and Biomass Valorization*, 4(2), 221-235.
- 49
50
51 755 Ministry of Finance of Vietnam, 2015. Financial resources for environmental protection for the
52
53 756 period 2011 - 2015, orientation for the period 2016 - 2020. The 4th National Environment
54
55 757 Conference, 2015.
- 56
57
58
59
60
61
62
63
64
65

1 758 Ministry of Natural Resources and Environment of Vietnam, 2016. Report on the current state of
2 759 the national environment for the period 2011 - 2015. Hanoi: Publishing House of Natural
3
4 760 Resources - Environment and Maps of Vietnam.
5
6
7 761 Ministry of Natural Resources and Environment of Vietnam, 2017. Report on the current state of
8
9 762 the national environment 2016. Hanoi: Publishing House of Natural Resources -
10
11 763 Environment and Map of Vietnam.
12
13
14 764 Ministry of Natural Resources and Environment of Vietnam, 2018. Report on the current state of
15
16 765 the national environment 2017. Hanoi: Publishing House of Natural Resources -
17
18 766 Environment and Maps of Vietnam.
19
20
21 767 Ministry of Natural Resources and Environment of Vietnam, 2020. Report on the state of the
22
23 768 national environment in 2019. Hanoi: Dan Tri Publishing House.
24
25
26 769 Moh, Y., 2017. Solid waste management transformation and future challenges of source separation
27
28 770 and recycling practice in Malaysia. Resources, Conservation and Recycling. 116, pp.1-14.
29
30
31 771 Morrison, M., Bennett, J., & Blamey, R., 1999. Valuing improved wetland quality using choice
32
33 772 modeling. Water resources research. 35(9), 2805-2814.
34
35
36 773 Nainggolan, D., Pedersen, A. B., Smed, S., Zemo, K. H., Hasler, B., & Termansen, M., 2019.
37
38 774 Consumers in a circular economy: economic analysis of household waste sorting
39
40 775 behaviour. Ecological Economics, 166, 106402.
41
42
43 776 Nainggolan, R., Perangin-angin, R., Simarmata, E., & Tarigan, A. F., 2019. Improved the
44
45 777 performance of the K-means cluster using the sum of squared error (SSE) optimized by
46
47 778 using the Elbow method. In Journal of Physics: Conference Series (Vol. 1361, No. 1, p.
48
49 779 012015). IOP Publishing.
50
51
52
53 780 Othman, J., 2007. Economic valuation of household preference for solid waste management in
54
55 781 Malaysia: A choice modeling approach. International Journal of Management Studies
56
57 782 (IJMS). 14(1), 189-212.
58
59
60
61
62
63
64
65

- 1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65
- 783 Othman, J., 2002. Household preferences for solid waste management in Malaysia. EEPSEA
784 research report series/IDRC. Regional Office for Southeast and East Asia, Economy and
785 Environment Program for Southeast Asia; no. 2002-RR8.
- 786 Padilla, A. J., & Trujillo, J. C., 2018. Waste disposal and households' heterogeneity. Identifying
787 factors shaping attitudes towards source-separated recycling in Bogotá, Colombia. *Waste
788 Management*, 74, 16-33.
- 789 Pearce, D., Howarth, A., 2000. Technical report on methodology: cost benefit analysis and policy
790 responses. RIVM report 481505020. National Institute of Public Health and the
791 Environment.
- 792 Pearce, D., Atkinson, G., & Mourato, S., 2006. Cost-benefit analysis and the environment: recent
793 developments. Organisation for Economic Co-operation and development.
- 794 Pek, C. K., & Othman, J., 2010. Household demand for solid waste disposal options in Malaysia.
- 795 Rahji, M. A. Y., & Oloruntoba, E. O., 2009. Determinants of households' willingness to pay for
796 private solid waste management services in Ibadan, Nigeria. *Waste management &
797 research*. 27(10), 961-965.
- 798 Ready, R. C., Buzby, J. C., & Hu, D., 1996. Differences between continuous and discrete
799 contingent value estimates. *Land Economics*. 397-411.
- 800 Rosenberger, R.S., Loomis, J.B., 2001. Benefit transfer of outdoor recreation use values: a
801 technical document supporting the Forest Service Strategic Plan (2000 revision). General
802 Technical Report-Rocky Mountain Research Station, USDA Forest Service.
- 803 Sakata, Y., 2007. A choice experiment of the residential preference of waste management
804 services–The example of Kagoshima city, Japan. *Waste management*. 27(5), 639-644.
- 805 Soderholm, P. (Ed.), 2013. *Environmental policy and household behaviour: sustainability and
806 everyday life*. Routledge.

1 807 Stevens, T. H., Belkner, R., Dennis, D., Kittredge, D., & Willis, C., 2000. Comparison of
2 808 contingent valuation and conjoint analysis in ecosystem management. *Ecological*
3
4 809 *Economics*. 32(1), 63-74.
5
6
7 810 Tarfasa, S., & Brouwer, R., 2018. Public preferences for improved urban waste management: a
8
9 811 choice experiment. *Environment and Development Economics*. 23(2), 184.
10
11
12 812 Thomas, C., & Sharp, V., 2013. Understanding the normalisation of recycling behaviour and its
13
14 813 implications for other pro-environmental behaviours: A review of social norms and
15
16 814 recycling. *Resources, Conservation and Recycling*. 79, 11-20.
17
18
19 815 Thurstone, L. L., 1927. A law of comparative judgment. *Psychological Review*. 34(4), 273.
20
21
22 816 Verplanken, B., & Aarts, H., 1999. Habit, attitude, and planned behaviour: is habit an empty
23
24 817 construct or an interesting case of goal-directed automaticity?. *European review of social*
25
26 818 *psychology*. 10(1), 101-134.
27
28
29 819 Webb, T. L., Sheeran, P., & Luszczynska, A., 2009. Planning to break unwanted habits: Habit
30
31 820 strength moderates implementation intention effects on behaviour change. *British Journal of*
32
33 821 *Social Psychology*. 48(3), 507-523.
34
35
36 822 World Bank, 2018. Report on solid and industrial hazardous waste management assessment
37
38 823 options and action areas to implement the national strategy. Hong Duc Publishing House.
39
40 824 <https://documents1.worldbank.org/curated/en/352371563196189492/pdf/Solid-and->
41
42 825 [industrial hazardous-waste-management-assessment-options-and-actions-areas.pdf](https://documents1.worldbank.org/curated/en/352371563196189492/pdf/Solid-and-industrial-hazardous-waste-management-assessment-options-and-actions-areas.pdf)
43
44
45
46 826 Whitehead, J. C., 2006. A practitioner's primer on the contingent valuation method. *Handbook on*
47
48 827 *contingent valuation*, 66-91.
49
50
51 828 Willis, K. G., & Garrod, G. D., 1999. Angling and recreation values of low-flow alleviation in
52
53 829 *rivers*. *Journal of environmental management*, 57(2), 71-83.
54
55
56
57
58
59
60
61
62
63
64
65

830 Yuan, Y., & Yabe, M., 2015. Residents' preferences for household kitchen waste source
1
2 831 separation services in Beijing: A choice experiment approach. International journal of
3
4 832 environmental research and public health. 12(1), 176-190.
5
6

7 833 Yusuf, S. A., Salimonu, K. K., & Ojo, O. T., 2007. Determinants of willingness to pay for
8
9 834 improved household solid waste management in Oyo State, Nigeria. Research journal of
10
11 835 applied sciences. 2(3), 233-239.
12
13

14 836
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65