

ACCEPTED VERSION

Declan R. Morris, Todd J. McWhorter, Wayne S. J. Boardman, Greg Simpson, Jeanette Wentzel, Jannie Coetzee, Ferreira Du Plessis, Yoshan Moodley

Gene flow connects key leopard (*Panthera pardus*) populations despite habitat fragmentation and persecution

Environmental Management, 2023; 71(2):260-273

© The Author(s), under exclusive licence to Springer Nature B.V. 2022. Springer Nature or its licensor (e.g. a society or other partner) holds exclusive rights to this article under a publishing agreement with the author(s) or other rightsholder(s); author self-archiving of the accepted manuscript version of this article is solely governed by the terms of such publishing agreement and applicable law.

This version of the article has been accepted for publication, after peer review (when applicable) and is subject to Springer Nature's [AM terms of use](#), but is not the Version of Record and does not reflect post-acceptance improvements, or any corrections. The Version of Record is available online at: <http://dx.doi.org/10.1007/s10531-022-02531-5>

PERMISSIONS

<https://www.springernature.com/gp/open-research/policies/journal-policies>

Self-archiving of papers published via the subscription route

Accepted Manuscript terms of use

Use of the AM is subject to an embargo period and our [AM terms of use](#), which permit users to view, print, copy, download and text and data-mine the content, for the purposes of academic research, subject always to the full conditions of use. Under no circumstances may the AM be shared or distributed under a Creative Commons, or other form of open access license, nor may it be reformatted or enhanced.

Authors should provide the following acknowledgement, and link from the accepted manuscript version to the URL of the published article on the journal's website

- *This version of the article has been accepted for publication, after peer review (when applicable) and is subject to Springer Nature's [AM terms of use](#), but is not the Version of Record and does not reflect post-acceptance improvements, or any corrections. The Version of Record is available online at: [http://dx.doi.org/\[insert DOI\]](http://dx.doi.org/[insert DOI])*

Embargoes and restrictions on self-archiving of the AM

For information about self-archiving of papers published via the subscription route (green open access), including embargo periods, please consult our table below:

Portfolio	Embargo length (articles)	Deposition of AM permitted in institutional or funder repository after embargo, and author's own personally maintained website*?
Nature Portfolio journals	6 months	Yes
Hybrid/transformational academic journals on nature.com platform	6 months	Yes
Springer hybrid/transformational and subscription journals	12 months	Yes
Palgrave Macmillan hybrid/transformational and subscription journals	12 months	Yes

7 June 2024

<http://hdl.handle.net/2440/141148>

1 **Gene flow connects key leopard (*Panthera pardus*) populations despite**
2 **habitat fragmentation and persecution**

3 Short Running Title: Leopard gene flow in fragmented landscapes

4 **Declan R. Morris¹, Todd J. McWhorter^{1*}, Wayne S.J. Boardman^{1*}, Greg Simpson², Jeanette**
5 **Wentzel³, Jannie Coetzee⁴, Ferreira Du Plessis⁴ & Yoshan Moodley⁵**

- 6 1. School of Animal and Veterinary Sciences, University of Adelaide, Roseworthy, South
7 Australia, 5371, Australia.
- 8 2. Faculty of Veterinary of Science, Department of Production Animal Studies,
9 University of Pretoria, Onderstepoort, Gauteng, 0110, South Africa.
- 10 3. Faculty of Veterinary of Science, Department of Veterinary Tropical Diseases,
11 University of Pretoria, Hans Hoheisen Wildlife Research Centre, 1380, South Africa.
- 12 4. Mpumalanga Tourism and Parks Agency, Nelspruit, Mpumalanga, 1200, South Africa.
- 13 5. Department of Zoology, University of Venda, Private Bag X5050, Thohoyandou 0950,
14 Republic of South Africa

15 **ORCID IDs -**

16 Declan Morris 0000-0003-3386-5214

17 Wayne Boardman [0000-0002-1746-0682](https://orcid.org/0000-0002-1746-0682)

18 Todd McWhorter 0000-0002-4746-4975

19 **Corresponding Author**

20 Declan Morris, School of Animal and Veterinary Sciences, University of Adelaide,
21 Roseworthy, South Australia, 5371, Australia.

22 declanmorris743@gmail.com

23

24 **Abstract**

25 The leopard (*Panthera pardus*) is facing the threat of continued population decline across its
26 range. In order to inform more effective conservation management programs, genetic
27 information is needed from leopard populations that persist in previously unstudied,
28 isolated and highly fragmented protected areas. The aim of this study was to explore the
29 population structure and genetic diversity of leopard populations across the Mpumalanga
30 province of South Africa. We collected a total of 33 leopard samples from four major
31 locations across a west to east transect across the province. We analysed 17 polymorphic
32 microsatellites and two regions of the mitochondrial genome (NADH-5 and Cytb) to
33 determine the genetic structure of the leopard population in the province. We also
34 calculated genetic diversity indices and explored gene flow in the region. We found that
35 while there is gene flow occurring across the province, there is still fine scale genetic
36 structuring of the population. We identified two major population units that we describe as
37 'West Mpumalanga' and 'East Mpumalanga'. Gene flow was moderate between the two
38 populations and we found very high genetic diversity levels compared to other leopard
39 populations previously studied in South Africa. From a conservation perspective, our results
40 show that gene flow is still occurring across seemingly isolated leopard populations that
41 exist in fragmented landscapes, highlighting the importance of all leopard populations in
42 South Africa. Management authorities need to focus conservation efforts on maintaining
43 corridors between regions that are suitable for leopard occupancy and work closely with
44 human settlements to minimise human-leopard conflicts.

45

46

47 **Keywords**

48 Leopard, gene flow, connectivity, Mpumalanga, population genetics, conservation

49 **Acknowledgments**

50 We would like to thank the Mpumalanga Tourism and Parks Agency for their support,
51 especially Gerrie Camacho, Delecia Gunn, Ernest Rohm, Gerhardus Sterk and Chris Hobkirk
52 their expertise input, assistance and time sacrificed for leopard captures. We would also like
53 to thank Jannie Du Bruyn, Dr. Hannes Botha, Tommie Steyn, Juan De Beer & Johan Eksteen
54 for their technical support and advice during the project. Thanks to Dr. Philip Stott for his
55 input into the project development. Thanks also to the students Herman Viviers, Tora-Lee
56 Axelson, Max Peters, Parris Jeffries, Marco Gouws and Olivia Rynders who assisted with field
57 work activities. This study was funded from PhD student operating funds available from the
58 School of Animal and Veterinary Science at The University of Adelaide.

59 **Declarations**

60 **Funding:** This study was funded from PhD student operating funds available from the School
61 of Animal and Veterinary Science at The University of Adelaide.

62 **Conflicts of Interest:** The authors declare no conflicts of interest

63 **Availability of data and material**

64 Curative plans for data accessibility will be as follows:

65 All mtDNA strands have been uploaded to Genebank under project no. XXX, accession
66 numbers x-y. Microsatellite data were uploaded to Dryad; There is no restrictions on data
67 availability.

68

69

70 **Author Contributions:** All authors had input to the overall project research design. Field
71 work research and sample collection was performed by D.R.M., J.C., G.S., F.P. & J.W.. J.W.
72 contributed genetic samples for analysis. Data analysis was performed by D.R.M. & Y.M..
73 The paper was written by D.R.M with large input from Y.M., T.J.M. & W.S.J.B.

74 **Ethics Approval:** Ethics approval was received from the University of Adelaide Animal Ethics
75 Committee (S-2016-023) and the permit to conduct research on Loskop Dam Nature
76 Reserve was given by the provincial municipality Mpumalanga Tourism and Parks Agency
77 (TS3/11).

78 **Consent for Publication:** All authors give their consent to the publication of this article

79

80

81

82

83

84

85

86

87

88 **Introduction**

89 Large carnivores play an important role in maintaining and regulating healthy ecosystems
90 (Miller et al. 2001; Ripple et al. 2014). They have direct effects on large herbivore species
91 through predation and meso-carnivores through intraguild competition (Ripple et al. 2014).
92 Currently, 64% of large carnivore species worldwide are threatened with extinction while
93 80% of species show declining population trends (Wolf and Ripple 2018). Large carnivores
94 are particularly susceptible to the threat of extinction as they generally persist at low
95 densities, have large home ranges and require large continuous areas of natural habitat to
96 hunt and reproduce (Cardillo et al. 2005). As the human population continues to grow,
97 natural habitats are being rapidly destroyed for new agricultural and urban developments.
98 This has led to the loss of both suitable habitat and widespread habitat fragmentation.
99 Habitat fragmentation impedes normal animal dispersal by introducing both physical (large
100 distances between suitable habitat, farms, fences etc.) and anthropogenic barriers (lethal
101 persecution / snares as a result from human wildlife conflicts) (Schlaepfer et al. 2018).
102 Animal dispersal is a critical factor in maintaining biodiversity as it is the process underlying
103 gene flow and genetic exchange between populations. Gene flow is an important
104 evolutionary process as it maintains or replenishes the genetic diversity of a population
105 (Lenormand 2002). Restricted gene flow results in population differentiation, with the
106 distribution of genetic variation becoming limited and compartmentalised across the
107 landscape (Moodley et al. 2017). If a population experiencing restricted gene flow is also
108 declining in size, detrimental genetic processes such as the loss of alleles due to genetic drift
109 and the accumulation of deleterious mutations through inbreeding can occur (Andersen et
110 al. 2004; Keyghobadi 2007). These processes will eventually lead to a net loss of genetic

111 diversity and decreased population viability due to a reduction in fitness and adaptive
112 potential (Keyghobadi 2007). Reduced genetic diversity can decrease a population's
113 resistance to disease (Spielman et al. 2004) and reduced fitness can manifest itself as
114 changes in birthweights, longevity and predation success rates (Keller and Waller 2002).
115 Population genetic studies examining the genetic structure of target species and gene flow
116 between identified populations have thus become a crucial element in the development and
117 implementation of successful conservation programs (Abdul-Muneer 2014; Kenney et al.
118 2014; Moodley et al. 2018). Defining the genetic structure of an animal population involves
119 determining (i) the number of existing subpopulations, (ii) the connectivity between these
120 subpopulations and (iii) the allele frequencies within these subpopulations (Chakraborty
121 1993; Mondol et al. 2013; Zanin et al. 2016). This information can be used to draw
122 boundaries and delineate evolutionarily significant or management units, their levels of
123 genetic diversity, degree of isolation or connectivity and inbreeding.

124 One country in which the genetics of large carnivores has been studied extensively is South
125 Africa. Although this African republic still harbours much of its large carnivore biodiversity,
126 its highly developed agricultural sector and growing human population mean that the
127 ranges of these species are mostly confined to government and private conservation areas.
128 It is therefore not surprising that species such as lion (*Panthera leo*), cheetah (*Acinonyx*
129 *jubatus*) and African wild dog (*Lycaon pictus*) have already started to exhibit the detrimental
130 genetic consequences of habitat fragmentation and subsequent isolation (Kotze et al. 2008;
131 Marsden et al. 2012; Miller et al. 2015). The South African leopard (*Panthera pardus pardus*)
132 is a large carnivore that is unique in that only 8.4% of its extant range exists inside of
133 protected areas (Jacobson et al. 2016). Due to their shy and elusive nature, leopards can live

134 in close proximity with humans and move undetected across large distances (Fattebert et al.
135 2013). Therefore, it is possible leopards do not display the same genetic patterns of the
136 other carnivores.

137 Several studies have explored the diversity and structure of both nuclear (microsatellite)
138 and mitochondrial DNA (mtDNA) in southern African leopard populations (Anco et al. 2018;
139 McManus et al. 2015; Ropiquet et al. 2015; Uphyrkina et al. 2001). Early mtDNA analysis
140 revealed two maternal lineages among leopards across Africa, with both lineages observed
141 in South Africa (Uphyrkina et al. 2001). Nuclear DNA studies using microsatellites reflect a
142 degree of isolation-by-distance (IBD) across southern Africa (Ropiquet et al. 2015), with local
143 structure observed between the Western and Eastern Cape Provinces (McManus et al.
144 2015). McManus et al. (2015) also estimated low to moderate gene flow connecting
145 subpopulations within the Western and Eastern Cape. Yet, despite their importance to the
146 conservation and management of leopards (Naude et al. 2020), gene flow and connectivity
147 between populations have not been extensively studied in South Africa.

148 The province harbouring one of the largest leopard populations in South Africa,
149 Mpumalanga, has not yet been studied in depth. Mpumalanga is located in north-eastern
150 South Africa and contains some of the largest wild areas in the country, including the Kruger
151 National Park, which also forms part of the Greater Limpopo Trans-frontier conservation
152 agreement. Its main distinguishing feature is a major west-east altitudinal drop as the
153 highveld plateau gives way to the low-lying plains (lowveld), via the Drakensburg mountain
154 escarpment. Mpumalanga thus contains three major biomes, highveld grassland (61% of
155 land area), lowveld savannah (39%) and escarpment forest (0.5%) (Ferrar and Lötter 2007).
156 Human land use across this province predominantly involves mining, manufacturing and

157 agriculture, and these industries have led to habitat fragmentation and the repurposing of
158 36% of the natural habitat (Ferrar and Lötter 2007). Detailed land use maps for Mpumalanga
159 have been published in Ferrar and Lötter (2007); Lötter (2015); Simpson et al. (2019).
160 Despite this, approximately 34% of Mpumalanga is still considered suitable habitat for
161 leopard occupancy (Swanepoel et al. 2013), with a leopard population estimated between
162 338 and 1,851 individuals (203 – 1,111 mature individuals) (Swanepoel et al. 2016;
163 Swanepoel et al. 2014). This combination of anthropogenic pressure and large numbers of
164 free-ranging large carnivores has brought humans and leopards into closer proximity,
165 resulting in a higher occurrence of human-leopard conflicts (Balme et al. 2010), with real or
166 perceived threats to livelihood (livestock and game) and human safety often resulting in
167 persecution killings of leopards.

168 Given the persistence of significant leopard numbers in this diverse and fragmented
169 landscape, one might expect reduced gene flow owing to either geographic barriers (e.g.
170 altitude or Isolation by Distance (IBD)) or anthropogenic factors, leading to increased
171 population structuring due to genetic drift. Understanding the genetic properties of the
172 current leopard population of Mpumalanga is essential for its conservation and
173 management. The aims of this study were therefore to determine: (i) whether Mpumalanga
174 leopards were genetically structured into subpopulations (ii) the genetic diversity of these
175 subpopulations and (iii) whether subpopulations were connected by contemporary gene
176 flow.

177

178 **Materials and Methods**

179 **Samples and study area**

180 In order to characterise leopard population structure, diversity and gene flow across
181 Mpumalanga's dramatic highveld-lowveld altitudinal gradient, 33 leopard samples were
182 collected from four sampling locations along a roughly west-east transect across the
183 northern part of the province (Figure 1). In the highveld region, samples were collected from
184 Loskop Dam Nature Reserve (LDNR) (n=13) and the Lydenburg district (n=3). LDNR has an
185 altitude that ranges between 1285 – 1406 m while Lydenburg sits at approximately 1414m.
186 In the lowveld region, we sampled at Andover Nature Reserve (ANR) (n=3) and Manyeleti
187 Nature Reserve (MNR) (n=12). While ANR (altitude 508 m) is fenced, nearby MNR (altitude
188 350-450m) is a 23,000 hectare reserve that forms part of the Greater Kruger National park
189 and is open to Kruger National Park (KNP).

190 All capture procedures were approved by the Animal Ethics Committees of the University of
191 Adelaide (S-2016-023) and The University of Pretoria (V115-16) and permission to conduct
192 the study was obtained from Mpumalanga Tourism and Parks Agency (MTPA, Project
193 Approval TS3/11). Thirteen LDNR samples were collected from captured animals using single
194 door cage traps which are activated by either a foot plate or a release mechanism placed on
195 bait. Traps were approximately 90 cm wide, 110 cm tall and 1200 cm long. Traps were set
196 up next to roads and game paths often utilised by leopards, based on the results of
197 extensive camera trapping that was simultaneously conducted on the reserve. A South
198 African Veterinary Council registered veterinarian anaesthetised the animals by darting
199 using a combination of zolazepam and tiletamine, at 1-3 mg/kg (Zoletil, Virbac, Centurion,
200 South Africa) and medetomidine at 0.05–0.09 mg/kg (Medetomidine, Kyron Laboratories,

201 Johannesburg, South Africa). Reversal of the anaesthesia was accomplished by
202 intramuscular injection of atipamezole at 0.25 – 0.45 mg/kg (Antisedan, Zoetis, Sandton,
203 South Africa) (Kock and Burroughs 2012).

204 Whole blood was collected in EDTA tubes and stored by both freezing at approximately -
205 20 °C and on Whatman FTA cards (Merk, Darmstadt, Germany). Three additional whole
206 blood samples from Lydenburg were collected opportunistically by the MTPA during
207 standard management practices. In addition, twelve blood samples from MNR and three
208 from ANR were collected as part of the annual census. Two skin samples were collected
209 from post-mortems on recently deceased leopards found on Manyeleti NR and two skin
210 samples were collected from confiscated leopard skins that were found in villages bordering
211 ANR and MNR. The exact origins of the confiscated skins are unknown, but are very likely to
212 have come from ANR, MNR or KNP.

213 **Molecular Methods**

214 DNA was extracted from the defrosted blood samples using a Quick-DNA Miniprep Plus Kit
215 D4068 (Zymo Research, CA, USA). Eighteen polymorphic microsatellites were selected for
216 analysis based on their use in previous leopard studies in South Africa (McManus et al. 2015;
217 Ropiquet et al. 2015; Uphyrkina et al. 2001). The markers [FCA008, FCA032, FCA075,
218 FCA077, FCA082, FCA085, FCA129, FCA133, FCA161, FCA191, FCA211, FCA224, FCA229,
219 FCA261, FCA310, FCA391, FCA441 & Y2F1-T3-4] were first isolated in the domestic cat (*Felis*
220 *catus*) and were all found to be polymorphic in leopards (Menotti-Raymond et al. 1999).
221 Microsatellites were amplified *via* polymerase chain reaction (PCR) using the Q5 Hot Start
222 High-Fidelity Kit (New England Biolabs, MA, USA) according to the manufacturer's
223 specifications. Reactions were contained in a total volume of 25 µL comprising 1.25 µL each

224 of forward and reverse primers, 1.0 μ L of extracted template DNA, 12.5 μ L master mix
225 (M0494S) and 9 μ L of nuclease free water (Amresco E476). Thermocycling included an initial
226 denaturation for 30 s at 98 $^{\circ}$ C followed by 35 cycles of denaturation/annealing/elongation
227 for 10 s at 98 $^{\circ}$ C, 30 s at 60 $^{\circ}$ C and 20 s at 72 $^{\circ}$ C followed by a final extension for 5 min at
228 72 $^{\circ}$ C. PCR products were genotyped on an ABI Prism 3500XL. Allele scoring was conducted
229 on SoftGenetic GeneMarker (v 2.6.7) and double checked by eye.

230 Additionally, two mitochondrial DNA markers were sequenced for all samples. NADH
231 dehydrogenase 5 (NADH-5) and Cytochrome-b (Cytb), were selected for amplification based
232 on their successful use in previous leopard studies (Ropiquet et al. 2015). NADH-5 was
233 amplified using the primers F/RL2 (Uphyrkina et al. 2001) and Cytb was amplified using the
234 primers Leo-F/Leo-R (Ropiquet et al. 2015). PCR was performed in 25 μ L reactions
235 comprising 0.5 μ L of each forward and reverse primer, 1.0 μ L of extracted DNA, 12.5 μ L
236 OneTaq Quick-Load 2X Master mix with buffer M0486 (New England Biolabs, MA, USA) and
237 10.5 μ L of nuclease free water (Amresco E476). Thermocycling included an initial
238 denaturation for 30 s at 94 $^{\circ}$ C followed by 35 cycles of denaturation/annealing/elongation
239 for 30 s at 94 $^{\circ}$ C, 30 s at 50 $^{\circ}$ C and 1 min at 68 $^{\circ}$ C followed by a final extension for 2 min at
240 68 $^{\circ}$ C. Amplicons were post-PCR purified with 2.5 μ L exonuclease I (Exo I) and 5 μ L shrimp
241 alkaline phosphatase (rSAP) (New England Biolabs, MA, USA) according to the
242 manufacturers instruction. The purified amplicons were then sequenced in both the forward
243 and reverse directions using the BIG Dye Kit (Applied Biosystems, Thermo Fisher Scientific,
244 CA, USA) and electrophoresed through an ABI PRISM3500xl Genetic Analyser using a 50cm
245 array and the POP-7 polymer (Applied Biosystems, Thermo Fisher Scientific, CA, USA).
246 Sequences were checked by eye, then forward and reverse sequences were aligned into
247 contigs, consensus bases were called and trimmed in BioEdit (v7.0.5.3) (Hall 1999). The final

248 NADH-5 alignment was 426bp in length (positions 12,632 – 13,058 on the complete mtDNA
249 genome of *Felis catus*, Accession number: U20753) (Lopez et al. 1996), while the cyt b
250 alignment was 1,137bp in length (positions 15,039 – 16,176 on the *Felis catus* mtgenome)

251 **Microsatellite structure**

252 Microsatellites were analysed in MICROCHECKER (v2.2.3) (Van Oosterhout et al. 2004) to
253 quality check the data and ensure that there were no genotyping errors, allele dropout or
254 null alleles present in the dataset. FSTAT (v2.9.4) (Goudet 1995) was used to test for linkage
255 disequilibrium and deviations from Hardy-Weinberg equilibrium were detected in
256 GENEALEX (6.503) (Peakall and Smouse 2012). The genetic structure of leopards across our
257 Mpumalanga transect was initially analysed using the program STRUCTURE (v.2.3.4)
258 (Pritchard et al. 2000). STRUCTURE uses a Bayesian approach to identify the most likely
259 number of populations (K) observed within a dataset by clustering multi-locus genotypes
260 together that share similar patterns of variation (Porrás-Hurtado et al. 2013). We modelled
261 up to 10 potential population clusters (K) with a Markov chain length of 1,000,000 iterations
262 and a burn-in period of 100,000. We used the admixture model with correlated allele
263 frequencies as the leopard populations across Mpumalanga could have been in recent
264 genetic contact and likely share common ancestors. The number of populations was
265 determined using the programs STRUCTURE HARVESTER (v.0.6.94) (Earl and Vonholdt 2012)
266 and CLUMPAK (Kopelman et al. 2015), following K -selection methods described by Evanno
267 et al. (2005). In addition, we used discriminate analysis of principal components (DAPC) to
268 infer genetic structure (Jombart et al. 2010). This model-free multivariate method does not
269 assume Hardy-Weinberg equilibrium nor linkage disequilibrium. The method was
270 implemented in the R-package 'ADEGENET' (Jombart 2008) for $K=1-10$. We also

271 implemented a spatial approach to explore genetic structure across the Mpumalanga
272 landscape using GENELAND (v4.0.7) (Guillot et al. 2005). Parameters in GENELAND were set
273 to model $K=1-10$, with 10 repetitions for each K , for 1,000,000 MCMC iterations. Finally, we
274 also tested the microsatellite data for patterns of IBD using a Mantel test of geographic and
275 genetic distance matrices, conducted in GENEALX.

276 **Mitochondrial DNA structure**

277 The maternal genetic structure among Mpumalanga leopards was also examined. Since DNA
278 sequence data for the same markers were also available for 16 samples from the nearby
279 Kruger National Park (Ropiquet et al. 2015), we downloaded these data (GenBank accession
280 NADH-5: JF70216-224, JF70234-236, JF70270-274; Cytb: JF720084-20092, JF720103-105,
281 JF720139-142) and included them in our mtDNA analyses. The leopard samples we
282 sequenced for these markers were uploaded to GenBank under the following accession
283 numbers (See Data Accessibility). Initially, mtDNA genetic structure was examined by
284 constructing minimum spanning haplotype networks from the concatenated mtDNA
285 alignments in POPART (v.3.2) (Leigh and Bryant 2015). To further determine genetic
286 structure, we reconstructed a phylogenetic tree using the program BEAST (Bayesian
287 Evolutionary Analysis by Sampling Trees) (v.2.5.2) (Bouckaert et al. 2014). Three outgroup
288 taxa were added to the sample set comprising three lion sequences (Genbank accession
289 numbers KP001498, KP001502 & KP001506) and Asiatic leopard subspecies *P.p fuscia* and
290 *P.p japonesis* (Genbank accession numbers EF199742, EF199743 & KJ866876). A nucleotide
291 substitution model for each partition (alignment) was selected by first analysing the
292 sequence data in JmodelTest (v2.1.10) (Darriba et al. 2012). The selected models were HKY +
293 G for both mitochondrial genes alignments based on BIC (Bayesian information criterion)

294 calculations. Heterogeneity in mutation rates was therefore modelled using a gamma
295 distribution with four bins. A Relaxed Clock Log Normal model was selected because its
296 posterior standard deviation for the mean clock rate did not include zero, implying that a
297 strict clock was not appropriate for this data set. A Markov chain of length 100,000,000 was
298 used to explore tree space, sampling a tree every 10,000 iterations, and discarding 15% as
299 burn-in. TRACER (v1.7.1) (Rambaut et al. 2018) was used to check that the Markov chain had
300 reached stationarity. Nodes that were not supported by a posterior probability > 50% were
301 collapsed using TreeGraph 2 (v.2.15.0) (Stover and Muller 2010).

302

303 **Genetic Diversity**

304 Genetic diversity for microsatellite loci was calculated using GENEALEX. Diversity was
305 examined using the average number of alleles per locus (A), observed heterozygosity (H_o),
306 expected heterozygosity (H_E) and the inbreeding coefficient (F_{IS}). Pairwise estimates of
307 fixation index (F_{ST}) were calculated between the defined population groups to infer the
308 degree of differentiation between them. MtDNA diversity indices included haplotype
309 diversity, nucleotide diversity and the number of polymorphic sites and were calculated in
310 DNASP (v.6.12.01) (Rozas et al. 2017). We also conducted neutrality tests, Tajima's D and
311 Fu's f_s statistic, to explore the likelihood of recent demographic changes.

312

313

314

315

316 **Gene flow**

317 Gene flow was inferred using microsatellite data. A rough measure of past levels of gene
318 flow was calculated from F_{ST} using the formula $(1/F_{ST}-1)/4$ (Wright 1984). Contemporary, and
319 bidirectional gene flow between structured populations was inferred using a Bayesian
320 approach implemented in BIMr (Bayesian Inference of Migration rates) (v1.0) (Faubet and
321 Gaggiotti 2008). We ran the program using the F-model (correlated allele frequencies) using
322 a Markov chain of 500,000 iterations and a burn-in period of 20,000 iterations. We ran five
323 Markov chains in total, checked that they had converged in TRACER, then averaged the
324 complementary gene flow between populations from each run as detailed in Epps et al.
325 (2013).

326

327

328 **Results**

329 Null alleles were flagged at three loci [FCA224, FCA075, Y2-F1-T34], however only Y2-F1-T34
330 contained more than 25% null alleles and was omitted from any further analysis
331 (Oosterhout > 0.25). Therefore, a total of 17 loci were included in all further analyses. No
332 loci were found to be statistically linked nor significantly deviated from Hardy-Weinberg
333 expectation ($p > 0.05$), after correction for multiple testing. MtDNA could not be amplified
334 from the two dried skin samples and thus, were not included in the mtDNA analysis.

335 **Microsatellite genetic structure**

336 The highest number of population clusters (Evanno et al. 2005) detected by STRUCTURE was
337 five ($K=5$), suggesting that biologically meaningful population structure could potentially
338 exist at $K \geq 5$ (Figure 2, S6). At $K=2$, the LDNR samples from the highveld separated from all
339 the others into a western Mpumalanga group, with the eastern Mpumalanga group
340 containing all other highveld and lowveld samples. At $K=3$, the three Lydenburg samples
341 from the highveld grouped together with one sample from ANR and another from MNR. At
342 $K=4$, the emergence of a 4th grouping comprised mostly of samples from MNR, but with one
343 sample from LDNR (green). At $K=5$, a further cluster was observed among the lowveld
344 samples from ANR and MNR. In general, however, the STRUCTURE analysis highlighted
345 several potentially admixed leopard genotypes at all locations as K increased (Figure 2, S6).

346 In contrast, model free DAPC clustering fitted the samples most optimally into two
347 populations, with individual grouping consistent with the STRUCTURE results of $K=2$ (Figure
348 S1-3). Our landscape analysis with GENELAND, which takes spatial information into account,
349 suggested a best fit into three geographic populations ($K=3$,) (Figure S4+S5, Table S1). The
350 only difference between this and $K=2$ was the separation of the three Lydenburg samples,

351 two ANR and three MNR samples into a third population, similar to the $K=3$ result from
352 STRUCTURE. The landscape analysis described a similar trend to both STRUCTURE and DAPC
353 analyses, in that the LDNR population was always partitioned separately from samples
354 collected in eastern Mpumalanga (Lydenburg, ANR, MNR). We also tested our microsatellite
355 data for patterns of IBD, and found only a weak correlation between geography and
356 genetics with a correlation co-efficient (R_{xy}) of 0.32, suggesting that IBD only accounts for
357 10% of the genetic structure observed among Mpumalanga leopards (Sample Size=33, No.
358 in Matrix = 528, P-value = 0.01) (Figure S7). We, hereafter, considered the Mpumalanga
359 landscape to be broadly divided into two main nuclear genetic populations inhabiting the
360 eastern and western parts of the province. We define population 1 as 'West Mpumalanga'
361 which consists of all thirteen LDNR samples and population 2 as 'East Mpumalanga' which
362 contains all samples collected from Lydenburg, ANR and MNR.

363 **Mitochondrial genetic structure**

364 The median spanning haplotype network of the concatenated NADH-5 and Cytb alignments
365 showed considerable maternal structure across Mpumalanga (Figure 3A). All haplotypes
366 were structured into two prominent haplogroups, which we call here, East and West
367 Mpumalanga. A Bayesian phylogenetic tree rooted with Asiatic leopard and lion confirmed
368 this structure (Figure 3B) and showed that while no sampling location constituted a
369 monophyletic clade, all but one Loskop sample occurred in the East Mpumalanga clade. The
370 East Mpumalanga clade is made up of all the samples from Lydenburg and the majority of
371 the lowveld samples (61.3%). The West Mpumalanga clade consists of the remaining 38.7%
372 lowveld samples and almost all of the samples from Loskop (92.3%).

373

374 **Genetic Diversity**

375 Genetic diversity indices were calculated for populations 'West Mpumalanga' and 'East
376 Mpumalanga' as defined by our genetic structure analyses. Nuclear genetic diversity was
377 reasonably high, with observed and expected heterozygosity > 0.7 for both West and East
378 Mpumalanga populations, as well as the entire data set, with mildly negative inbreeding
379 coefficients (Table 1). West Mpumalanga displayed slightly fewer alleles per locus than East
380 Mpumalanga.

381 We also explored genetic diversity using the mitochondrial data following the same
382 population definitions (West and East Mpumalanga as above). However, for mtDNA
383 diversity, the addition of the sixteen KNP samples were included in the 'East Mpumalanga'
384 population grouping. The East Mpumalanga population was more diverse, containing 16
385 haplotypes compared to only six within the West Mpumalanga population. Similarly,
386 haplotype diversity, nucleotide diversity and the number of polymorphic sites were also
387 higher in the East Mpumalanga clade (Table 2). Tajimas D returned low and non-significant
388 values.

389 **Gene flow**

390 Both methods for calculating nuclear gene flow showed genetic connectivity between the
391 West Mpumalanga and East Mpumalanga leopard populations. Using pairwise F_{ST} , which is a
392 rough estimate of long-term rate of migration, the rate of exchange was approximately
393 three individuals per generation (3.09). However, although bidirectional Bayesian estimates
394 of gene flow were similarly high, there was a marked difference in directionality. Migration
395 from western to eastern Mpumalanga (3.3%) was almost three times lower than the rate of

396 migration in the opposite direction (9.7%), suggesting a greater net movement of leopards
397 from east to west across Mpumalanga (Table S2).

398

399 **Discussion**

400 This is the first analysis of leopard population genetic structure occurring across the
401 ecologically diverse landscape of Mpumalanga Province, South Africa. Our results show that
402 while there is fine scale structuring occurring in the province, these subpopulations remain
403 connected as there is gene flow occurring between them. This supports the findings of other
404 studies that conclude the leopard population within Southern Africa comprises a continuous
405 metapopulation but is subject to genetic structuring at a regional scale (McManus et al.
406 2015; Ropiquet et al. 2015; Spong et al. 2000). Leopard genetic structuring has also been
407 found to occur at the wider African continental scale (Anco et al. 2018).

408 **The structure of leopard populations across Mpumalanga**

409 Our results provide compelling evidence for the structuring of leopard genetic variation
410 across the landscape of northern Mpumalanga. In all analyses, nuclear genetic markers
411 consistently partitioned the leopards of LDNR as separate to the rest of the province. The
412 pairwise F_{st} between the two structured populations was moderate (0.056), implying some
413 degree of differentiation, but with significant gene flow connecting the two populations.
414 However, the partitioning of other subpopulations at greater values of K , only within East
415 Mpumalanga, suggests that this region could harbour other pockets of leopard diversity.
416 The fact that three highveld samples taken at Lydenburg always separated at $K=3$ also

417 implies some degree of distinctiveness around the escarpment region, at the interface
418 between highveld and lowveld.

419 This nuclear population structure was further supported by even more pronounced
420 maternal structuring of mtDNA genes. Both analyses of mtDNA structure partitioned the
421 majority of LDNR samples (92.3%) into their own West Mpumalanga clade separate to the
422 ANR and Lydenburg samples. A major difference however, between the mtDNA and
423 microsatellite structure results, is that leopard samples from MNR and KNP shared
424 haplotypes with both East and West Mpumalanga clades. This observation is consistent with
425 other mtDNA studies of southern African leopards (Anco et al. 2018; Ropiquet et al. 2015).

426 Potential explanations for this observed structure of leopard populations across
427 Mpumalanga include the large altitudinal gradient separating the two halves of the
428 province. However, as samples from Lydenburg (highveld) consistently partitioned within
429 the East Mpumalanga clade and not with LDNR (also highveld) samples in the West
430 Mpumalanga clade, it is unlikely that genetic differences are attributable to altitude. The
431 observation of lowveld samples from ANR and MNR portioning together with distant
432 Lydenburg at *K3* and within the East Mpumalanga mtDNA clade, strongly implies that the
433 altitudinal difference between high- and lowveld does not pose a barrier to gene flow in
434 leopards. Another explanation could be that localised dispersal and genetic drift resulted in
435 a gradient of isolation by distance across the region. In this situation, patchy sampling across
436 the landscape could give the illusion of two distinct population groups. However, we found
437 that only 10% of the observed structure can be attributed to IBD. This finding, together with
438 the high number of lowveld haplotypes from KNP and MNR clustering in the West
439 Mpumalanga clade suggest that both populations could have diverged in historical isolation

440 and have since come into more recent secondary genetic contact via gene flow across the
441 Drakensberg escarpment. However, a larger nuclear and mitochondrial data set of South
442 African and other African leopards would be required to test this hypothesis.

443 **Connectivity of leopard populations across Mpumalanga**

444 While mtDNA haplotype sharing could hint at population connectivity of leopard
445 populations between western and eastern Mpumalanga, it is also possible that the lack of
446 mtDNA regional monophyly stems from the presence of ancestral haplotypes, which are
447 thus still present in both West and East clades because of incomplete lineage sorting.
448 Microsatellites, on the other hand, are rapidly evolving and therefore more conducive for
449 analyses of recent gene flow and contemporary population connectivity (Feulner et al. 2004;
450 Teske et al. 2018). We found that, despite habitat fragmentation due to human activities,
451 leopard populations in Mpumalanga remain connected to each other by contemporary gene
452 flow. Leopard generation time is approximately 6-7 yrs (Balme et al. 2013), and at a
453 minimum, one migrant between each population every generation is required order to
454 maintain genetic diversity (Wang 2004). In the case of naturally occurring populations,
455 which are subject to various environmental conditions and species-specific factors, up to ten
456 migrants per generation has been suggested as a requirement to maintain maximal levels of
457 genetic diversity (Mills and Allendorf 1996). Here, we show that more gene flow (9.7%)
458 occurred from East Mpumalanga to West Mpumalanga than in the reverse direction (3.3%).
459 A net influx of contemporary genetic diversity into West Mpumalanga from East
460 Mpumalanga could be due to the high density of leopard occurring in the greater Kruger
461 region and the distribution of leopard-suitable habitat in the province. The high density of
462 leopards in KNP, which was estimated to be 12.7 - 30.9 (Bailey 1993; Maputla et al. 2013)

463 individuals per 100 km², may be what drives this east to west connectivity, as leopard
464 density and competition is significantly lower at 8 individuals per 100 km² in Loskop Dam
465 Nature Reserve (Morris et al. 2021). These microsatellite-derived gene flow estimates
466 therefore complement the mtDNA results, demonstrating that non-monophyly in both West
467 and East Mpumalanga mtDNA clades is likely due to gene flow, rather than incomplete
468 lineage sorting.

469 **Genetic diversity**

470 The heterozygosity levels we recorded for Mpumalanga leopard populations are the highest
471 reported in South Africa to date. All values, whether for the entire province or West or East
472 Mpumalanga populations, returned observed heterozygosity values of between 0.72-0.78,
473 which are significantly higher ($H_0 \pm 2SE$) than reported values from the Western Cape
474 (0.624-0.657), Eastern Cape (0.646-0.657) and KwaZulu-Natal (0.638-0.660) or Mozambique
475 (0.687) (McManus et al. 2015; Ropiquet et al. 2015). The heterozygosity of Mpumalanga
476 leopards was thus more in line with East African leopards (H_0 Tanzania = 0.77) (Spong et al.
477 2000), than other southern African leopards. Although these values were compared against
478 different, and in some cases significantly fewer loci, we are still confident that our data set
479 of 18 polymorphic microsatellites is comparable, since loci were chosen specifically to
480 overlap with all previous studies. The average number of microsatellite alleles was higher
481 among leopards in East Mpumalanga, and consistent with the equivalent statistic in our
482 mtDNA data set (number of haplotypes, Table 2), which was much higher in the East clade.
483 The East Mpumalanga clade also displays more polymorphic sites and higher haplotype
484 diversity compared to the West Mpumalanga clade. These similarities in patterns of genetic
485 diversity between microsatellite and mtDNA data sets also strongly hints at the evolutionary

486 equivalence of the observed microsatellite populations and mtDNA clades. Whether genetic
487 structuring and patterns of diversity are similar in other parts of South Africa where
488 leopards are still free-ranging (such as Limpopo, North-West and Northern Cape Provinces)
489 is presently unknown, however, we show here that even opportunistically collected samples
490 that are patchily distributed across the landscape can shed considerable light on the
491 structure and diversity of leopard populations.

492 Mpumalanga leopards display relatively high genetic diversity in part due to the gene flow
493 that is occurring between East and West Mpumalanga clades, but there is also likely to be
494 gene flow occurring between neighbouring provinces. LDNR is closely located to large
495 portions of leopard suitable habitat located in southern Limpopo that stretches and
496 connects much of the Limpopo province (Swanepoel et al. 2013). The Eastern Mpumalanga
497 population unit is highly connected to other regions in the lowveld such as KwaZulu-Natal,
498 KNP, Mozambique and Limpopo, through conservation agreements such as the Greater
499 Limpopo trans frontier conservation agreement. A detailed map of this suitable leopard
500 habitat across South Africa is available to view in Swanepoel et al. (2013). We suggest that
501 while there is still gene flow occurring across the east-west Mpumalanga transect, this
502 connectivity is probably not the only contribution to the maintenance of high genetic
503 diversity of Mpumalanga's leopards.

504 **Conservation outlook and conclusion**

505 While genetic diversity within the sampled Mpumalanga populations remains relatively
506 high, it can also decrease rapidly within generations if gene flow is impeded (Kotze et al.
507 2019). Anthropogenic landscape use and human-leopard conflicts have the potential to
508 impede gene flow. Currently only 14.8% of the land cover in Mpumalanga is protected

509 (Ferrar and Lötter 2007). Additionally, 19.3% of the Mpumalanga land cover is made up of
510 cultivated areas for farming purposes, 1.0% used for mining and 2.8% in urban areas (Lötter
511 2015). The land along the northern transverse of the province that links our studies' sample
512 sites has been classified as predominantly grazing and poorly adapted cultivated land
513 (Simpson et al. 2019). There is also highly adapted cultivated land found directly
514 neighbouring LDNR and the again to the east just before the KNP region (Simpson et al.
515 2019). This brings leopards into direct contact with humans where conflicts, such as
516 livestock predation, often occur. In South Africa, leopards accounted for between 40-89% of
517 livestock attacks across the country (Constant et al. 2015; Thorn et al. 2012). Landowners,
518 therefore, often view leopards negatively (Grey et al. 2017), with 98% of interview
519 participants from communities that border the KNP perceiving predators as a major threat
520 to their livestock (Legendijk and Gusset 2008), and 67% of farmers have used lethal methods
521 to resolve conflicts with leopards (Thorn et al. 2012). It is estimated that a maximum of 169
522 leopards are removed from the Mpumalanga province annually either from retaliatory
523 killings or translocation of damage-causing individuals (Swanepoel et al. 2014).

524 Anthropogenic leopard mortalities are also attributed to motor vehicle accidents, legal
525 hunts and snares outside of protected areas (Swanepoel et al. 2015). In order to conserve
526 or potentially increase gene flow throughout Mpumalanga, naturally occurring leopard
527 dispersal corridors in Mpumalanga and to neighbouring provinces need to be identified and
528 preserved. On a national scale, stringent regulation of the trophy hunting industry and the
529 implementation of non-lethal control regulations for damage causing individuals, to
530 maintain or improve enhance diversity, should be investigated.

531 In order to implement effective leopard conservation management on a national scale,
532 genetic data for other unsampled parts of South Africa that have been identified as

533 important leopard habitat regions (i.e. Limpopo, Northern Cape and North West provinces)
534 (Swanepoel et al. 2013) are required. Furthermore, as each local study continues to conduct
535 genetic research in isolation, microsatellite genetic data sets are often not directly
536 comparable due to differences in markers used and the continued practice of scoring alleles
537 by fragment size rather than repeat number.

538 We have also identified that these previously unstudied leopard populations contain a high
539 level of genetic diversity and remain interconnected to neighbouring regions and reserves
540 despite being surrounded by a high density of human settlements. This finding helps
541 highlight, that even seemingly isolated leopard populations persisting in fragmented
542 landscapes, are important to the overall conservation management of the species. These
543 findings also have wider implications for other carnivore and large mammal species across
544 Mpumalanga as it shows despite the challenges, animals are still moving between protected
545 areas of the region.

546

547 **References**

- 548 Abdul-Muneer PM (2014) Application of Microsatellite Markers in Conservation Genetics and
549 Fisheries Management: Recent Advances in Population Structure Analysis and Conservation
550 Strategies. *Genetics Research International* 2014 doi:10.1155/2014/691759
- 551 Anco C, Kolokotronis SO, Henschel P, Cunningham SW, Amato G, Hekkala E (2018) Historical
552 mitochondrial diversity in African leopards (*Panthera pardus*) revealed by archival museum
553 specimens. *Mitochondrial DNA Part A* 29:455-473. doi:10.1080/24701394.2017.1307973
- 554 Andersen LW, Fog K, Damgaard C (2004) Habitat fragmentation causes bottlenecks and inbreeding in
555 the European tree frog (*Hyla arborea*). *Proc R Soc B-Biol Sci* 271:1293-1302.
556 doi:10.1098/rspb.2004.2720
- 557 Bailey TN (1993) *The African leopard : ecology and behavior of a solitary felid*. Columbia University
558 Press, New York
- 559 Balme GA, Batchelor A, Britz ND et al (2013) Reproductive success of female leopards *Panthera*
560 *pardus*: the importance of top-down processes. *Mammal Rev* 43:221-237.
561 doi:10.1111/j.1365-2907.2012.00219.x
- 562 Balme GA, Slotow R, Hunter LTB (2010) Edge effects and the impact of non-protected areas in
563 carnivore conservation: leopards in the Phinda-Mkhuze Complex, South Africa. *Anim*
564 *Conserv* 13:315-323. doi:10.1111/j.1469-1795.2009.00342.x

565 Bouckaert R, Heled J, Kuhnert D et al (2014) BEAST 2: A Software Platform for Bayesian Evolutionary
566 Analysis. *PLoS Comput Biol* 10:6. doi:10.1371/journal.pcbi.1003537

567 Cardillo M, Mace GM, Jones KE et al (2005) Multiple causes of high extinction risk in large mammal
568 species. *Science* 309:1239-1241. doi:10.1126/science.1116030

569 Chakraborty R (1993) Analysis of Genetic Structure of Populations: Meaning, Methods, and
570 Implications. In: Majumder PP (ed) *Human Population Genetics: A Centennial Tribute to J. B.*
571 *S. Haldane*. Springer US, Boston, MA, pp 189-206. doi:10.1007/978-1-4615-2970-5_14

572 Constant NL, Bell S, Hill RA (2015) The impacts, characterisation and management of human-leopard
573 conflict in a multi-use land system in South Africa. *Biodivers Conserv* 24:2967-2989.
574 doi:10.1007/s10531-015-0989-2

575 Darriba D, Taboada GL, Doallo R, Posada D (2012) jModelTest 2: more models, new heuristics and
576 parallel computing. *Nat Methods* 9:772-772. doi:10.1038/nmeth.2109

577 Earl DA, Vonholdt BM (2012) STRUCTURE HARVESTER: a website and program for visualizing
578 STRUCTURE output and implementing the Evanno method. *Conservation Genetics Resources*
579 4:359-361. doi:10.1007/s12686-011-9548-7

580 Epps CW, Castillo JA, Schmidt-Küntzel A, du Preez P, Stuart-Hill G, Jago M, Naidoo R (2013)
581 Contrasting Historical and Recent Gene Flow among African Buffalo Herds in the Caprivi Strip
582 of Namibia. *J Hered* 104:172-181. doi:10.1093/jhered/ess142

583 Evanno G, Regnaut S, Goudet J (2005) Detecting the number of clusters of individuals using the
584 software STRUCTURE: a simulation study. *Mol Ecol* 14:2611-2620. doi:10.1111/j.1365-
585 294X.2005.02553.x

586 Fattebert J, Dickerson T, Balme G, Slotow R, Hunter L (2013) Long-distance natal dispersal in leopard
587 reveals potential for a three-country metapopulation. *South Afr J Wildl Res* 43:61-67.
588 doi:10.3957/056.043.0108

589 Faubet P, Gaggiotti OE (2008) A new Bayesian method to identify the environmental factors that
590 influence recent migration. *Genetics* 178:1491-1504. doi:10.1534/genetics.107.082560

591 Ferrar AA, Lötter MC (2007) *Mpumalanga Biodiversity Conservation Plan Handbook*. Mpumalanga
592 Tourism and Parks Agency, Nelspruit

593 Feulner PGD, Bielfeldt W, Zachos FE, Bradvarovic J, Eckert I, Hartl GB (2004) Mitochondrial DNA and
594 microsatellite analyses of the genetic status of the presumed subspecies *Cervus elaphus*
595 *montanus* (Carpathian red deer). *Heredity* 93:299-306. doi:10.1038/sj.hdy.6800504

596 Goudet J (1995) FSTAT (Version 1.2): A computer program to calculate F-statistics. *J Hered* 86:485-
597 486. doi:10.1093/oxfordjournals.jhered.a111627

598 Grey JNC, Bell S, Hill RA (2017) Leopard diets and landowner perceptions of human wildlife conflict in
599 the Soutpansberg Mountains, South Africa. *J Nat Conserv* 37:56-65.
600 doi:10.1016/j.jnc.2017.03.002

601 Guillot G, Mortier F, Estoup A (2005) GENELAND: a computer package for landscape genetics.
602 *Molecular Ecology Notes* 5:712-715. doi:10.1111/j.1471-8286.2005.01031.x

603 Hall TA (1999) BioEdit: a user-friendly biological sequence alignment editor and analysis program for
604 Windows 95/98/NT. *Nucleic acids symposium series* 41:95-98.

605 Jacobson AP, Gerngross P, Lemeris JR et al (2016) Leopard (*Panthera pardus*) status, distribution, and
606 the research efforts across its range. *PeerJ* 4:28. doi:10.7717/peerj.1974

607 Jombart T (2008) adegenet: a R package for the multivariate analysis of genetic markers.
608 *Bioinformatics* 24:1403-1405. doi:10.1093/bioinformatics/btn129

609 Jombart T, Devillard S, Balloux F (2010) Discriminant analysis of principal components: a new
610 method for the analysis of genetically structured populations. *Bmc Genetics* 11
611 doi:10.1186/1471-2156-11-94

612 Keller LF, Waller DM (2002) Inbreeding effects in wild populations. *Trends in Ecology & Evolution*
613 17:230-241. doi:10.1016/s0169-5347(02)02489-8

614 Kenney J, Allendorf FW, McDougal C, Smith JL (2014) How much gene flow is needed to avoid
615 inbreeding depression in wild tiger populations? *Proceedings of the Royal Society B:*
616 *Biological Sciences* 281:20133337. doi:doi.org/10.1098/rspb.2013.3337

617 Keyghobadi N (2007) The genetic implications of habitat fragmentation for animals. *Can J Zool*
618 85:1049-1064. doi:10.1139/z07-095

619 Kock MD, Burroughs REJ (2012) *Chemical and physical restraint of wild animals : a training and field*
620 *manual for African species. Second edition. edn. IWVS, Greyton*

621 Kopelman NM, Mayzel J, Jakobsson M, Rosenberg NA, Mayrose I (2015) Clumpak: a program for
622 identifying clustering modes and packaging population structure inferences across K.
623 *Molecular Ecology Resources* 15:1179-1191. doi:10.1111/1755-0998.12387

624 Kotze A, Ehlers K, Cilliers DC, Grobler J (2008) The power of resolution of microsatellite markers and
625 assignment tests to determine the geographic origin of cheetah (*Acinonyx jubatus*) in
626 Southern Africa. *Mammalian Biology - MAMM BIOL* 73:457-462.
627 doi:10.1016/j.mambio.2007.10.011

628 Kotze A, Smith RM, Moodley Y et al (2019) Lessons for conservation management: Monitoring
629 temporal changes in genetic diversity of Cape mountain zebra (*Equus zebra zebra*). *PLoS One*
630 14:14. doi:10.1371/journal.pone.0220331

631 Legendijk DDG, Gusset M (2008) Human-Carnivore Coexistence on Communal Land Bordering the
632 Greater Kruger Area, South Africa. *Environ Manage* 42:971-976. doi:10.1007/s00267-008-
633 9204-5

634 Leigh JW, Bryant D (2015) POPART: full-feature software for haplotype network construction.
635 *Methods in Ecology and Evolution* 6:1110-1116. doi:10.1111/2041-210x.12410

636 Lenormand T (2002) Gene flow and the limits to natural selection. *Trends in Ecology & Evolution*
637 17:183-189. doi:10.1016/s0169-5347(02)02497-7

638 Lopez JV, Cevario S, O'Brien SJ (1996) Complete nucleotide sequences of the domestic cat (*Felis*
639 *catus*) mitochondrial genome and a transposed mtDNA tandem repeat (Numt) in the nuclear
640 genome. *Genomics* 33:229-246. doi:10.1006/geno.1996.0188

641 Lötter MC (2015) *Technical Report for the Mpumalanga Biodiversity Sector Plan – MBSP.*
642 *Mpumalanga Tourism and Parks Agency, Nelspruit*

643 Maputla NW, Chimimba CT, Ferreira SM (2013) Calibrating a camera trap-based biased mark-
644 recapture sampling design to survey the leopard population in the N'wanetsi concession,
645 Kruger National Park, South Africa. *Afr J Ecol* 51:422-430. doi:10.1111/aje.12047

646 Marsden CD, Woodroffe R, Mills MGL et al (2012) Spatial and temporal patterns of neutral and
647 adaptive genetic variation in the endangered African wild dog (*Lycaon pictus*). *Mol Ecol*
648 21:1379-1393. doi:10.1111/j.1365-294X.2012.05477.x

649 McManus JS, Dalton DL, Kotze A, Smuts B, Dickman A, Marshal JP, Keith M (2015) Gene flow and
650 population structure of a solitary top carnivore in a human-dominated landscape. *Ecol Evol*
651 5:335-344. doi:10.1002/ece3.1322

652 Menotti-Raymond M, David VA, Lyons LA, Schaffer AA, Tomlin JF, Hutton MK, O'Brien SJ (1999) A
653 genetic linkage map of microsatellites in the domestic cat (*Felis catus*). *Genomics* 57:9-23.
654 doi:10.1006/geno.1999.5743

655 Miller B, Dugelby B, Foreman D, del Rio CM, Noss R, Phillips M (2001) The importance of large
656 carnivores to healthy ecosystems. *Endangered species update* 18:202.

657 Miller SM, Harper CK, Bloomer P, Hofmeyr J, Funston PJ (2015) Fenced and Fragmented:
658 Conservation Value of Managed Metapopulations. *PLoS One* 10:e0144605.
659 doi:10.1371/journal.pone.0144605

660 Mills LS, Allendorf FW (1996) The one-migrant-per-generation rule in conservation and
661 management. *Conserv Biol* 10:1509-1518. doi:10.1046/j.1523-1739.1996.10061509.x

662 Mondol S, Bruford MW, Ramakrishnan U (2013) Demographic loss, genetic structure and the
663 conservation implications for Indian tigers. *Proc R Soc B-Biol Sci* 280:10.
664 doi:10.1098/rspb.2013.0496

665 Moodley Y, Russo I-RM, Dalton DL et al (2017) Extinctions, genetic erosion and conservation options
666 for the black rhinoceros (*Diceros bicornis*). *Scientific Reports* 7:41417.
667 doi:10.1038/srep41417

668 Moodley Y, Russo IRM, Robovsky J et al (2018) Contrasting evolutionary history, anthropogenic
669 declines and genetic contact in the northern and southern white rhinoceros (*Ceratotherium*
670 *simum*). *Proc R Soc B-Biol Sci* 285 doi:10.1098/rspb.2018.1567

671 Morris DR, Boardman WS, Swanepoel LH, Simpson G, Coetsee J, Camacho GJ, McWhorter TJ (2021)
672 Population density estimate of leopards (*Panthera pardus*) in north-western Mpumalanga,
673 South Africa, determined using spatially explicit capture–recapture methods. *Mamm Biol*:1-
674 11.

675 Naude VN, Balme GA, O'Riain J, Hunter LTB, Fattedbert J, Dickerson T, Bishop JM (2020) Unsustainable
676 anthropogenic mortality disrupts natal dispersal and promotes inbreeding in leopards. *Ecol*
677 *Evol* 10:3605-3619. doi:10.1002/ece3.6089

678 Peakall R, Smouse PE (2012) GenAlEx 6.5: genetic analysis in Excel. Population genetic software for
679 teaching and research-an update. *Bioinformatics* 28:2537-2539.
680 doi:10.1093/bioinformatics/bts460

681 Porras-Hurtado L, Ruiz Y, Santos C, Phillips C, Carracedo A, Lareu MV (2013) An overview of
682 STRUCTURE: applications, parameter settings, and supporting software. *Front Genet* 4:98-
683 98. doi:10.3389/fgene.2013.00098

684 Pritchard JK, Stephens M, Donnelly P (2000) Inference of population structure using multilocus
685 genotype data. *Genetics* 155:945-959. doi:doi.org/10.1093/genetics/155.2.945

686 Rambaut A, Drummond AJ, Xie D, Baele G, Suchard MA (2018) Posterior Summarization in Bayesian
687 Phylogenetics Using Tracer 1.7. *Syst Biol* 67:901-904. doi:10.1093/sysbio/syy032

688 Ripple WJ, Estes JA, Beschta RL et al (2014) Status and Ecological Effects of the World's Largest
689 Carnivores. *Science* 343:151-162. doi:10.1126/science.1241484

690 Ropiquet A, Knight AT, Born C et al (2015) Implications of spatial genetic patterns for conserving
691 African leopards. *C R Biol* 338:728-737. doi:10.1016/j.crv.2015.06.019

692 Rozas J, Ferrer-Mata A, Sanchez-DelBarrio JC, Guirao-Rico S, Librado P, Ramos-Onsins SE, Sanchez-
693 Gracia A (2017) DnaSP 6: DNA Sequence Polymorphism Analysis of Large Data Sets.
694 *Molecular Biology and Evolution* 34:3299-3302. doi:10.1093/molbev/msx248

695 Schlaepfer DR, Braschler B, Rusterholz HP, Baur B (2018) Genetic effects of anthropogenic habitat
696 fragmentation on remnant animal and plant populations: a meta-analysis. *Ecosphere* 9
697 doi:10.1002/ecs2.2488

698 Simpson GB, Badenhorst J, Jewitt GPW, Berchner M, Davies E (2019) Competition for Land: The
699 Water-Energy-Food Nexus and Coal Mining in Mpumalanga Province, South Africa. *Frontiers*
700 *in Environmental Science* 7 doi:10.3389/fenvs.2019.00086

701 Spielman D, Brook BW, Briscoe DA, Frankham R (2004) Does inbreeding and loss of genetic diversity
702 decrease disease resistance? *Conserv Genet* 5:439-448.
703 doi:10.1023/B:COGE.0000041030.76598.cd

704 Spong G, Johansson M, Bjorklund M (2000) High genetic variation in leopards indicates large and
705 long-term stable effective population size. *Mol Ecol* 9:1773-1782. doi:10.1046/j.1365-
706 294x.2000.01067.x

707 Stover BC, Muller KF (2010) TreeGraph 2: Combining and visualizing evidence from different
708 phylogenetic analyses. *Bmc Bioinformatics* 11 doi:10.1186/1471-2105-11-7

709 Swanepoel LH, Balme G, Williams S et al (2016) A conservation assessment of *Panthera pardus*. In: In
710 Child MF RL, Do Linh San E, Raimondo D, Davies-Mostert HT, editors (ed) *The Red List of*
711 *Mammals of South Africa, Swaziland and Lesotho*. South African National Biodiversity
712 *Institute and Endangered Wildlife Trust, South Africa.*, pp 1-13

713 Swanepoel LH, Lindsey P, Somers MJ, van Hoven W, Dalerum F (2013) Extent and fragmentation of
714 suitable leopard habitat in South Africa. *Anim Conserv* 16:41-50. doi:10.1111/j.1469-
715 1795.2012.00566.x

716 Swanepoel LH, Lindsey P, Somers MJ, Van Hoven W, Dalerum F (2014) The relative importance of
717 trophy harvest and retaliatory killing of large carnivores: South African leopards as a case
718 study. *South Afr J Wildl Res* 44:115-134. doi:10.3957/056.044.0210
719 Swanepoel LH, Somers MJ, van Hoven W et al (2015) Survival rates and causes of mortality of
720 leopards *Panthera pardus* in southern Africa. *Oryx* 49:595-603.
721 doi:10.1017/s0030605313001282
722 Teske PR, Golla TR, Sandoval-Castillo J et al (2018) Mitochondrial DNA is unsuitable to test for
723 isolation by distance. *Scientific Reports* 8:8448. doi:10.1038/s41598-018-25138-9
724 Thorn M, Green M, Dalerum F, Bateman PW, Scott DM (2012) What drives human-carnivore conflict
725 in the North West Province of South Africa? *Biol Conserv* 150:23-32.
726 doi:10.1016/j.biocon.2012.02.017
727 Uphyrkina O, Johnson WE, Quigley H, Miquelle D, Marker L, Bush M, O'Brien SJ (2001) Phylogenetics,
728 genome diversity and origin of modern leopard, *Panthera pardus*. *Mol Ecol* 10:2617-2633.
729 doi:10.1046/j.0962-1083.2001.01350.x
730 Van Oosterhout C, Hutchinson WF, Wills DPM, Shipley P (2004) MICRO-CHECKER: software for
731 identifying and correcting genotyping errors in microsatellite data. *Molecular Ecology Notes*
732 4:535-538. doi:10.1111/j.1471-8286.2004.00684.x
733 Wang JL (2004) Application of the one-migrant-per-generation rule to conservation and
734 management. *Conserv Biol* 18:332-343. doi:10.1111/j.1523-1739.2004.00440.x
735 Wolf C, Ripple WJ (2018) Rewilding the world's large carnivores. *Royal Society Open Science* 5
736 doi:10.1098/rsos.172235
737 Wright S (1984) *Evolution and the genetics of populations vol 3*. Univeristy of Chicago Press, Chicago
738 Zanin M, Adrados B, González N et al (2016) Gene flow and genetic structure of the puma and jaguar
739 in Mexico. *European journal of wildlife research* 62:461-469. doi:0.1007/s10344-016-1019-8

740

741

742

743

744

745

746

747

748

749

750

751 **Figure Legends**

752 **Figure 1:** The study area of Mpumalanga Province, South Africa. A. Mpumalanga Province
753 relative to the rest of South Africa. B. Sampling areas along a west-east gradient across
754 Mpumalanga's highveld and Lowveld regions. Blue shading represents the four sampling
755 locations: Loskop Dam Nature Reserve (1), Lydenburg (2), Andover Nature Reserve (3) and
756 Manyeleti Game Reserve (4).

757

758 **Figure 2:** The structure of genetic variation among leopards in a west-east transect across
759 Mpumalanga Province, South Africa. Each genotype is represented by a single line in the
760 plot and the contributions of different populations (*K*) are colour-coded. Models for
761 population clustering range from 2 to 5 (*K*2-*K*5). Regions where samples were obtained are
762 Loskop Dam Nature Reserve, Lydenburg (LYD), Andover Nature Reserve (AND), Manyeleti
763 Game Reserve (MNR) and confiscated leopard skin samples that were obtained from villages
764 located on the border of the greater Kruger region (SK). A consensus of three methods
765 suggested the greatest separation is between West Mpumalanga and East Mpumalanga
766 nuclear genetic populations.

767

768 **Figure 3:** Maternal genetic structure of leopards in Mpumalanga Province, South Africa **(A)**
769 Median spanning haplotype network based on NADH dehydrogenase 5 (NADH-5) and
770 Cytochrome-b (Cytb) mtDNA sequences. Two major clades are identified in West
771 Mpumalanga and East Mpumalanga. **(B)** Phylogeny showing the structure leopard mtDNA
772 haplogroups. The tree was constructed in BEAST. Nodes with <50% posterior probability
773 were collapsed.. The samples were coloured-coded according to sampled populations: Blue
774 – Loskop Dam Nature Reserve, Yellow – Lydenburg, Green – Andover Nature Reserve,
775 Purple – Manyeleti Game Reserve, Red – Kruger National Park. Two major clades were
776 identified, East Mpumalanga and West Mpumalanga.

777

778

779

780

781

782

783

784

785 **Tables – Table 1**

	n	H_o ± SEM	H_e ± SEM	F_{IS}	A ± SEM	PA	F_{ST}
Total	33	0.755 ± 0.025	0.735 ± 0.015	-0.027	6.44 ± 0.28	4.76	0.056
West Mpumalanga	13	0.787 ± 0.038	0.718 ± 0.023	-0.096	5.59 ± 0.38	3.15	-
East Mpumalanga	20	0.722 ± 0.031	0.751 ± 0.020	0.039	7.29 ± 0.31	5.80	-

786

787 **Table 1:** Genetic diversity of leopard populations in Mpumalanga, South Africa. Diversity
788 indices are given for the entire provincial sample (total) and for the two major
789 subpopulation samples identified in analyses of genetic structure (East and West
790 Mpumalanga). n=number of samples, H_o = observed heterozygosity, H_e = expected
791 heterozygosity, F_{IS} = inbreeding coefficient, A = number of alleles averaged for all loci, PA =
792 private alleles, F_{ST} = pairwise population comparison.

793

794

795

796

797

798

799

800

	No of samples	No. of haplo types	Haplotype diversity (hd)	Nucleotide diversity (Pi)	Polymorphic sites	Tajima's D	Fu's Fs
West Mpumalanga	13	6	0.782	0.00765	37	-0.010	4.332
East Mpumalanga	34	17	0.930	0.01073	44	1.596	1.430
Total	47	23	0.948	0.1266	52	2.301	0.919

801 **Table 2**

802

803

804 **Table 2:** Table of mtDNA genetic diversity statistics for the concatenated NADH-5 and Cytb
805 genes in leopard populations in Mpumalanga Province, South Africa. Diversity statistics are
806 given for the entire provincial population (total) and for the two major subpopulations
807 indented (East and West Mpumalanga). Diversity analyses performed are hd = haplotype
808 diversity, Pi = nucleotide diversity and the tests of neutrality Tajima's D and Fu's Fs.

809

810

811